

## Durham Research Online

---

### Deposited in DRO:

02 November 2018

### Version of attached file:

Accepted Version

### Peer-review status of attached file:

Peer-reviewed

### Citation for published item:

Masiero, M. and Franceschinis, C. and Mattea, S. and Thiene, M., D. and Pettenella, D. and Scarpa, R. (2018) 'Ecosystem services' values and improved revenue collection for regional protected areas.', *Ecosystem services*, 34 (A). pp. 136-153.

### Further information on publisher's website:

<https://doi.org/10.1016/j.ecoser.2018.10.012>

### Publisher's copyright statement:

© 2018 This manuscript version is made available under the CC-BY-NC-ND 4.0 license  
<http://creativecommons.org/licenses/by-nc-nd/4.0/>

### Additional information:

## Use policy

---

The full-text may be used and/or reproduced, and given to third parties in any format or medium, without prior permission or charge, for personal research or study, educational, or not-for-profit purposes provided that:

- a full bibliographic reference is made to the original source
- a [link](#) is made to the metadata record in DRO
- the full-text is not changed in any way

The full-text must not be sold in any format or medium without the formal permission of the copyright holders.

Please consult the [full DRO policy](#) for further details.

# Ecosystem services' values and improved revenue collection for regional protected areas

Mauro Masiero<sup>a</sup>, Cristiano Franceschinis<sup>a</sup>, Stefania Mattea<sup>b</sup>, Mara Thiene<sup>a</sup>, Davide Pettenella<sup>a</sup>, and Riccardo Scarpa<sup>b,c,d,#</sup>

<sup>a</sup> Land, Environment, Agriculture and Forestry Department - University of Padova  
Viale dell'Università, 16 - 35020 Legnaro (PD) – Italy  
davide.pettenella@unipd.it - mara.thiene@unipd.it - mauro.masiero@unipd.it -  
cristiano.franceschinis@unipd.it

<sup>b</sup> Waikato Management School, Economics Department - University of Waikato  
Hillcrest Road, Hamilton 3240 – New Zealand  
rscarpa@waikato.ac.nz - stefania.mattea@gmail.com

<sup>c</sup> Durham University Business School, Durham University, U.K.

<sup>d</sup> Department of Business Economics, University of Verona, Italy

# Corresponding author:

**Riccardo Scarpa** e-mail riccardo.scarpa@durham.ac.uk - tel. +44-(0)191-3347253 fax +44-(0)191-3345249

## Abstract

The management of conservation areas is a costly enterprise, especially vulnerable to budget cutting when austerity measures are being considered. Optimal spatial taxation dictates that tax-payers contribute proportionally to the benefits they receive. This paper provides a framework to derive spatially varied benefit estimates for ecosystem services produced in Natura 2000 protected areas of Lombardy (Italy). These may be used as a framework for spatially optimised taxation to improve the efficiency of public funding. In the process we used non-market valuation techniques, as well as benefit functions' transfer.

## Highlights

- A framework for the economic valuation of ecosystem services from Natura 2000 sites georeferenced across a wider region
- Analysis relies on choice experiments, benefit functions' transfer and mapping
- Average WTP values per person range between €2.28/year (slope protection) and €24.75 (carbon)

- Increases in ecosystem services supply correlate with increases in WTP value estimates but with significant differences depending on the ES taken into consideration

## **Keywords**

Natura 2000; protected areas; ecosystem services; choice experiment; benefit transfer; Lombardy.

## **1. Introduction and policy background**

Protected areas are aimed to conserve ecosystem integrity, safeguard ecological assets and maintain ecosystem services (ES) (Eastwood *et al.*, 2016). However, their public management is a costly enterprise, vulnerable to budget cutting when austerity measures are being considered. In order to improve management efficiency, specific spatially-targeted policies can be developed, to avoid that lack of spatial differentiation in the targeting mechanisms leads to efficiency losses (Wünscher *et al.*, 2008). Among spatially targeted policies, financing via spatially differentiated taxes represents one option (Pirard, 2012). The design of such policies, however, requires specific information on the locations of economic benefits generated by ES as provided by geographically specific protected areas. Optimal spatial taxation dictates that tax payers ought to contribute proportionally to the benefits they receive.

This study provides spatially varied benefit estimates for ES produced by Natura 2000 protected areas in the region of Lombardy (Italy). Estimates are developed through stated preferences methods (choice experiments) and may be used as both a framework for spatially optimised taxation and to improve the efficiency in collecting public funds.

The literature on spatial explicit willingness to pay (WTP) values in stated preferences research is quite vast and it addresses the topic from different angles. This includes, among others, literature on distance decay effect on benefits and the role of substitute sites (Smith, 1993; Bateman *et al.*, 2006; Schaafsma *et al.*, 2013), spatial patterns (Johnston and Duke, 2009; Johnston and Ramachandran, 2013; Johnston *et al.*, 2015; Holland and Johnston, 2017) and solutions for inferring how WTP values vary on maps, for example by means of interpolation methods and simple prediction (Campbell *et al.*, 2009; Czajkowski *et al.*, 2016; Sagebiel *et al.*, 2017). However the literature has paid limited attention to the challenges posed by spatial-specific assessment of ES (Zulian *et al.*, 2018)

70 to support policies and management practices in protected areas, including Natura 2000  
71 sites.

72 The concept of ES is over one century old (Marsh, 1864), however only in the last thirty  
73 years (Gómez-Baggethun *et al.*, 2010) has the scientific community forced the attention of  
74 policy makers to focus on the role that ES play in support of human activities. Interest in  
75 the issue has quickly grown from both public and private sectors (Ruckelshaus *et al.*,  
76 2015). With the publication of the Millennium Ecosystem Assessment (MEA, 2003) ES  
77 gained momentum within the international policy agenda as a way to improve the  
78 effectiveness of biodiversity-protection policies, thus encouraging research efforts (Fisher  
79 *et al.*, 2009). While at the beginning MEA did not pay much attention to the economics of  
80 ES (TEEB, 2010), in recent years interest in the investigation of their economic value has  
81 grown. Such interest is mostly motivated by the need to develop economic incentives for  
82 self-sustaining conservation activities (Jack *et al.*, 2008), through the creation of missing  
83 markets and the implementation of Payment for Ecosystem Services (PES) schemes to  
84 collect the necessary funds (Wunder, 2005).

85 ES have become a centrepiece of the European Union (EU) Biodiversity Strategy and their  
86 economic valuation can contribute to a better-informed decision-making (Maes *et al.*,  
87 2012; Schägner *et al.*, 2013). In its resolution of 12<sup>th</sup> December 2013 on green  
88 infrastructure, the EU Parliament emphasised the need to strengthen capacity and  
89 knowledge in relation to the mapping and assessment of specific ecosystems and their  
90 services.

91 Natura 2000 (COM, 2011) is a EU-wide network of nature protection areas. Such network  
92 was designated under the 1979 EU Birds (79/409/EEC, replaced by 2009/147/EC) and the  
93 1992 EU Habitats (92/43/EEC) Directives to ensure the conservation of valuable and  
94 threatened species/habitats. It consists of 27,308 terrestrial and marine Sites of  
95 Community Importance (SCI) and Special Protection Areas (SPAs) stretching over 100  
96 million ha (i.e. roughly 18% of EU territory) (European Commission, 2017). Areas in the  
97 Natura 2000 network are a cornerstone of the goals to be reached by the EU Biodiversity  
98 Strategy by the year 2020. Yet, to date, the potential of the ES concept to add value to  
99 current conservation approaches remains insufficiently explored (Harrison *et al.*, 2010; Ziv  
100 *et al.*, 2017). In particular, more socio-economic research on the effects of Natura 2000 is  
101 needed (Popescu *et al.*, 2014) given the current dearth of quantitative studies, especially  
102 of economic valuations of the benefits produced by the network (Gantioler *et al.*, 2014).  
103 Existing studies are very context-specific and have either sub-national or local scale

104 (Bastian, 2013; Gibson *et al.*, 2004; Chuan-Zhong *et al.*, 2004; Hoyos *et al.*, 2012). Some  
105 build on scaling-up of or extrapolating from available estimates (ten Brink *et al.*, 2013). In  
106 different EU countries, Schulp *et al.* (2014) highlighted substantial gaps between ES  
107 assessment and mapping exercises, due to the lack of both a systematic method of  
108 primary data collection and consistency across methodological approaches. This paper  
109 reports a study intended to fill this gap focussing on the most industrialised region of Italy:  
110 Lombardy.

111

112 In Italy, information gaps are exacerbated by the fact that it is one of the most bio-diverse  
113 countries in Europe (UNEP-WCMC, 2004), where the SCI and SPA network covers about  
114 one-fifth of the land (European Commission, 2017). Anthropogenic pressure on natural  
115 resources and local ecosystems is high, while at the same time human wellbeing is  
116 strongly reliant on them (MELS, 2013). This paradox is particularly evident in Lombardy. It  
117 is the Italian region hosting the highest number of Natura 2000 sites (242) and the most  
118 populated national region (about 10 million inhabitants, 16.5% of the total country  
119 population) and the second in terms of population density (414 inhabitants/km<sup>2</sup>) (Istat,  
120 2015). It is also the region with the highest rates of land use given up to urbanization  
121 (Ispra, 2015), a process seriously encroaching on ES from natural ecosystems (Turbè *et al.*,  
122 2010). It hosts more than one-fourth of the Italian industrial activities classified as  
123 hazardous for environmental resources and human health (Ispra, 2013). Environmental  
124 regulations (2015 National Budget Law) stress the need to both understand the economics  
125 of ES as well as to develop a system to account for their values and market provision.  
126 Although Italy has been included within a number of studies regarding Natura 2000, the  
127 identification and valuation of ES within the network has just been introduced through a  
128 couple of recent EU Life+ projects (Gestire<sup>1</sup> and Making Good Natura<sup>2</sup>) and addressed by  
129 publications developed within them (e.g. Schirpke *et al.*, 2014; Schirpke *et al.*, 2017;  
130 Schirpke *et al.*, 2018).

131 Economic valuation of ES is still the subject of a lively debate within the scientific  
132 community (Gómez-Baggethun and Ruiz Pérez, 2011). Influential researchers emphasise  
133 how knowledge gaps in this field may affect the capacity to inform policy (Ruckelshaus *et al.*,  
134 2015), thus potentially providing scope to improve the efficiency of natural resource

---

<sup>1</sup> [www.naturachevale.it/en/](http://www.naturachevale.it/en/)

<sup>2</sup> [www.lifemgn-serviziecosistemici.eu/EN/home/Pages/default.aspx](http://www.lifemgn-serviziecosistemici.eu/EN/home/Pages/default.aspx)

135 management and nature conservation (Pagiola *et al.*, 2004; Heal *et al.*, 2005; Silvis and  
136 van der Heide, 2013).

137 The present study aims to:

- 138 • develop a methodological approach for the economic assessment of the main ES  
139 provided by Natura 2000 network in Lombardy;
- 140 • test the approach with regards to a selection of non-marketed ES provided by two  
141 geographically separated protected areas within Lombardy;
- 142 • inform policy makers and set up guidelines for future periodical data collection and  
143 systematic accounting of ES values.

144 The study is novel since so far no economic assessment of ES provided by protected  
145 areas at the regional scale has been performed in Italy. Only few of them have been  
146 performed within EU (Christie and Rayment, 2012, Bateman *et al.*, 2013). The study  
147 adopts a valuable and innovative 6-steps methodological approach (see paragraph 2) the  
148 outcomes of which can contribute to inform future policies in this sector, by providing  
149 valuable inputs for decision-makers.

## 151 **2. A framework for Natura 2000 ES assessment in Lombardy**

152 The study builds on the six methodological steps adopted for the aims of Action 5 of  
153 Gestire project and presented in detail in Pettenella *et al.* (2016), i.e.:

- 154 (1) **literature review** of economic assessments of ES and Natura 2000 sites in Lombardy;
- 155 (2) **identification of the three main potential ES** provided by each of the 242 Natura  
156 2000 sites at regional scale. This included (i) an extensive analysis of the most recent (i.e.  
157 October 2014) official Standard Data Forms<sup>3</sup> for each site, as made available by the  
158 Ministry of Environment, Land and Sea, and (ii) identification and assessment of main  
159 potential ES per site based on a scoring system<sup>4</sup>. We revised assessments and scoring  
160 systems adopted in similar studies (Bastian, 2013; Schirpke *et al.*, 2013) to link single  
161 habitat types to potential ES production. The most represented among ES categories—  
162 according to the Common International Classification of Ecosystem Services (CICES)  
163 version 4.3<sup>5</sup> (CICES, 2017)—are regulating (47%) and cultural ES (39%), while

---

<sup>3</sup> Standard Data Form have standardised structure and fields according to Commission Implementing Decision of 11 July 2011 concerning a site information format for Natura 2000 sites (notified under document C(2011) 4892)

<sup>4</sup> The following scores were used as a reference (Bastian, 2013; Schirpke *et al.*, 2013): 0 null ES potentiality, 1 low ES potentiality, 2, average ES potentiality, and 3 high ES potentiality.

<sup>5</sup> While other ES classification systems -such as for example the one developed by the Millennium Ecosystem Assessment- consider four ES categories (i.e. provisioning, regulating, cultural and supporting

164 provisioning ES are less frequent (14%). The three most relevant single ES cover about  
165 fifty percent of total ES potential production: C1, aesthetic value (21%), R9, biodiversity  
166 (17%) and C2, tourism and recreation (11%);

167 (3) **economic assessment of selected ES with a functioning market**, as produced by  
168 the Natura 2000 network in Lombardy. These included: fodder, timber, non-timber forest  
169 products (NTFPs), water provision, and carbon sequestration. Estimates were obtained  
170 making references to market prices or by means of different market value-based  
171 approaches (transformation costs, substitution costs, etc.) depending on the ES;

172 (4) **discrete choice experiment** (henceforth DCE, McFadden, 1973; Louviere, 1991;  
173 Boxall *et al.*, 1996) to estimate the marginal willingness-to-pay (WTP) for improving the  
174 quality of a selection of ES, treated as attributes in the DCE, and identified on the basis of  
175 step two. Two Natura 2000 sites were selected in Lombardy: Ticino and Adamello  
176 Regional Parks (see 2.1 below). They were chosen in order to cover both lowland (Ticino)  
177 and mountain (Adamello) areas within the region, which produce different sets of ES;

178 (5) Step five involved the estimation of **benefit function transfer (BFT)** to infer values of  
179 ES from study sites (for which we had sample observations) to policy sites (i.e. all  
180 municipalities within Lombardy without sample observations);

181 (6) The final step of the method involved **drawing conclusions and identifying future**  
182 **research needs**.

183 The core of the present study was in steps four and five, for which details are provided in  
184 Section 3 (Theory and Methods) below.

185

## 186 2.1 Study area

187 The 242 Natura 2000 sites in Lombardy cover two biogeographical regions (Figure 1):  
188 Alpine and Continental. Their combined area is about 372,000 ha, or about one-sixth of  
189 the region. These sites host 56 different habitats, 12 of which are considered of priority  
190 relevance according to EU Directives, as they are home to a variety of protected species:  
191 82 bird species, 83 other animal species (i.e. mammals, fish, invertebrates and  
192 amphibians) and 27 plant species (Regione Lombardia, 2018).

---

ES) the CICES focuses on the first three ES categories. It does not explicitly include the supporting ES because they are treated as part of the underlying structures, processes and functions that characterise ecosystems. These ES are indirectly consumed or used and they may simultaneously facilitate the output of many 'final outputs', therefore they were thought to be best dealt within environmental accounts through other ways. For further information see: <https://cices.eu/cices-structure/>.

193 The DCEs survey focused on the ES generated by two regional parks within Lombardy  
 194 (Figure 2): (i) **Adamello Regional Park** (henceforth Adamello RP) representing the Alpine  
 195 biogeography and (ii) **Ticino Regional Park** (henceforth Ticino RP) representing the  
 196 Continental biogeography. Adamello RP was created in 1983 and stretches over 51,000  
 197 ha in the North-East of Lombardy (Brescia province) at an altitude ranging from 390 to  
 198 3,591 mt above sea level (asl). The Adamello RP neighbours with two Italian parks and  
 199 one Swiss: altogether they form the largest continuous protected area within the Alps.  
 200 The Adamello RP includes 14 SCIs, covering around 5,550 ha, and 1 SPA (4,974 ha). It  
 201 also hosts part of the largest Italian glacier (Adamello glacier) and relevant prehistoric rock  
 202 and cave paintings dating back to the Iron Age and included within a UNESCO World  
 203 Heritage Site since 1979.  
 204 The Ticino RP, created in 1974, is the oldest regional park in Italy. It includes the Lombard  
 205 part of the Ticino River Valley and covers about 91,800 ha of lowland areas (56-427 m asl)  
 206 along the Ticino River. The RP includes municipal areas of all 47 municipalities located  
 207 along the river within the provinces of cities of Varese, Milan and Pavia. Ticino RP hosts  
 208 15 SCIs (10,971 ha) and 1 SPA (21,722 ha).

209

### 210 **3. Material and methods**

211 The ES under scrutiny, as produced by these protected areas, possess neither a proper  
 212 market nor a related market for a weakly complementary good. Economic valuation must  
 213 hence require non-market methods. To develop a value framework for the application of  
 214 any non-market valuation method, it is necessary to hinge the practice on a theory of value  
 215 consistent with individual utility theory. In our case the objective was to estimate the  
 216 economic value to the Lombardy residents derived from changes in the flows of selected  
 217 ES, as generated by the two Natura 2000 areas of Adamello and Ticino RPs.

218 Since utility is an ordinal concept, only utility changes can be associated with economic  
 219 values by using the equivalence principle between utility states. Let  $ES_0$  be the *status quo*  
 220 flow of ES (e.g. Alpine meadows left unmanaged – low endemic flora) and  $ES_1$  be the  
 221 proposed change brought about by a policy action that modifies such ES flow (e.g. 200 ha  
 222 of managed Alpine meadows – higher endemic flora). Then, the economic value to the  
 223 individual Lombardy resident derived from the proposed policy change is defined by the  
 224 compensating variation (CV) equivalence formula:

$$225 \quad U(ES_0; Y) = U(ES_1; Y + \Delta Y) = U(ES_1; Y + CV) \quad (1)$$



226 Where  $U(.)$  is the individual utility function,  $Y$  is income and  $\Delta Y$  defines the income change  
 227 necessary to offset the variation in ES. The correct welfare measure change is  $\Delta Y = CV$  and  
 228 its sign goes in the opposite direction of the perceived utility change without  
 229 compensation: an improvement in ES flow from the status-quo generates a negative CV,  
 230 as income level needs to be lowered to equalize utilities in the two endowment states (i.e.  
 231 a payment is due) (Freeman III *et al.*, 2014).

232 Random utility modelling of discrete choice responses collected in choice experiments  
 233 allows researchers to estimate the stochastic utility functions of the population from a  
 234 sample of respondents. Because of the obvious variation of preferences for ES across  
 235 residents, models with taste heterogeneity need to be fitted to the DCE data. We use a  
 236 finite mixture (Scarpa *et al.*, 2000; Boxall and Adamowicz, 2002; Scarpa and Thiene,  
 237 2005; Thiene *et al.*, 2015; Morey and Thiene, 2017) formulation of the mixed logit category  
 238 of models (see also 3.1 below), also known as Latent Class model (LCM). With such  
 239 estimates in hand, one can derive estimates of welfare change for specific policies  
 240 affecting the various ES subject to evaluation, using the equivalence above. Because of  
 241 the panel nature of the DCEs and the use of models with taste heterogeneity, such  
 242 estimates can be computed at the individual level using Bayes's theorem and observed  
 243 choice data (Train, 2003; Greene, Hensher and Rose, 2005; Scarpa *et al.*, 2008; Thiene *et al.*,  
 244 2013; Sarrias and Daziano 2018). Such estimates are then geo-referenced to the  
 245 municipalities sampled in the survey and value maps of ES values are obtained. These are  
 246 used to describe the spatial variation of economic values over the region.

247 Once the sample estimates for all marginal WTP changes are obtained they can be used,  
 248 in conjunction with socio-economic covariates and geographical data, to estimate separate  
 249 benefit functions for each marginal change and each ES. These functions describe how  
 250 estimates of value vary across residential locations and individuals in Lombardy. For  
 251 example, those residing far away from RPs might have lower values for ES improvements  
 252 everything else equal. Alternatively, those living in urban areas might have higher values  
 253 than those living in rural areas, because of the relative paucity of substitutes for ES in  
 254 urban areas. Such benefit functions are used to infer values in areas not covered by our  
 255 survey sampling by using the well-established technique of BFT (Loomis *et al.*, 1995;  
 256 Downing and Ozuna, 1996; Kirchhoff *et al.*, 1997; Bergstrom and Civita, 1999; Smith *et al.*,  
 257 2002; Vázquez-Polo *et al.*, 2002; Moeltner *et al.*, 2007; Johnston and Moeltner, 2014;  
 258 Moeltner and Rosenberger, 2014; Johnston *et al.*, 2015). This amounted to predicting

average economic values for ES changes for residents of municipalities not represented in our DCE sample, the “policy” sites (1018 for Adamello RP and 1004 for Ticino RP), using the values from the sampled municipalities, the “study” sites (505 for Adamello RP and 519 for Ticino RP). Predictions were obtained by using benefit functions based on determinants selected on the basis of prediction performance criteria (see 3.2 below). Maps with estimated or predicted values for ES can then be readily produced to illustrate to policy makers the distribution of values of each potential policy or combination of policies. These can be used to develop a revenue collection mechanism in which local taxes (regional rates) are spatially varied to match the spatial pattern of benefits as enjoyed by residents, delivering one of the principles of optimal tax theory and residential location theory. Incidentally, this could constitute a serious incentive for value revelation in public surveys as they would be perceived as highly consequential (Vossler *et al.*, 2012).

271

### 272 **3.1 Choice experiments**

After selecting a specific sub-set of ES for both Adamello RP and Ticino RP, two different online questionnaires were developed with their respective DCE surveys. Each of these were completed by about 1,500 respondents. Both samples were identified with the support of a company specialised in providing representative panels for on-line surveys: they included visitors and non-visitors, residing in Lombardy, aged 18-65, and were stratified according to selected socio-economic characteristics, as well as to the distance of their place of residence from the sites (5 zones). Each survey used a separate set of five policy dependent ES attributes and different policy-achievable levels were identified for each of them. The fifth was a proposed local annual tax increase, used as a payment vehicle, planned over a 5-year period and earmarked for expenditures necessary to deliver the quality improvement of the ES (Tables 1 and 2). Attributes and levels were discussed and agreed with the management staff of the two parks. Additional questions—based on a Likert-scale approach—were used to detect protest responses. This allowed those respondents who stated they were unwilling to pay additional taxes to elaborate on the reasons for such choice.

In order to reduce the risk of respondents ignoring one or more attributes included in the CE (attribute non-attendance), choice tasks were preceded by an introductory session aimed (among other things) at testing respondents’ familiarity with issues addressed by the survey. This session also provided basic information on the surveyed areas (including maps and pictures) and on topics covered by the choice experiment. Each respondent

293 was presented with 12 choice tasks and in each they were asked to select their most  
294 preferred policy scenario between three alternatives (examples of choice tasks are  
295 reported in the Appendix together with an example of the online survey). Having a  
296 relatively high number of choice tasks per respondent allowed us to increase the number  
297 of observations for choice models estimation. Fatigue effects which may arise during the  
298 sequence are a possible drawback of our design, as such effects could increase the  
299 degree of randomness of choices and/or increase the probability of adoption of heuristic  
300 strategies (Caussade *et al.*, 2005; Hensher, 2006). However, no relevant fatigue related  
301 issue emerged from the pre-test of the survey, so we leaned towards the certainty of  
302 having more data over the risk of a decrease in their accuracy. This choice was also  
303 corroborated by studies which found little to no evidence of fatigue effects (e.g. Carlsson  
304 *et al.*, 2012; Czajkowski *et al.*, 2014). The use of sequences longer than 10 choice tasks is  
305 also advisable to avoid bias in individual averages of marginal WTP estimates (Sarrias and  
306 Daziano 2018).

307 Opt-out options were not included as we wanted respondents to express a preference  
308 among possible ES improvement scenarios, given the relevance of such information for  
309 the authorities in charge of parks management. While this choice may have some  
310 drawback, the effects of the inclusion of the opt-out option are still debated (Veldwijk *et al.*,  
311 2014; Campbell and Erdem, 2018). A *status quo* alternative was not included to avoid any  
312 possible *status-quo* bias. A total number of 120 choice tasks were developed through an  
313 experimental design obtained with a dedicated software (Ngene by Choicemetrics, 2014).  
314 Results from preliminary pilot studies (about 30 per study-site) were used to design the  
315 surveys through a Bayesian efficient D-error minimizing design approach (Scarpa *et al.*,  
316 2007).

317 Data collected were used to estimate random utility models. Estimates were obtained for  
318 Multinomial Logit (MNL) and Latent Class Models (LCM), the latter models account for  
319 taste differences across respondents. We decided to explore taste heterogeneity by  
320 means of LCM rather than a Random Parameters Logit Model (RPL) because we found  
321 LCM to perform better on our data in preliminary analysis. Furthermore, LCM can be  
322 preferable to RPL when transferring results to policy makers (Sagebiel, 2017).

323 Choice models were used to derive individual marginal WTPs ( $WTP_m$ ). MNL models were  
324 estimated through NLOGIT version 5.0 software, while LCM through Latent Gold Choice  
325 version 4.5. Based on individual WTPs, average WTPs per municipality were computed  
326 and mapped via ArcGIS for municipalities covered by the two surveys.

327 In choice experiments, the sequence of individuals' choices is modelled as a function of  
 328 the attributes using Random Utility Theory (Luce, 1959; McFadden, 1973). According to  
 329 the Random Utility Theory, for an individual  $n$  facing a set of  $J$  alternatives, denoted by  
 330  $j=1,\dots,J$ , the utility of choosing the alternative  $i$  is a function of the  $K$  characteristics of the  
 331 alternative  $i$ . Utility functions are composed of a systematic part  $V_{ni}$  and a random part  
 332  $\varepsilon_i$  standing for all unobserved variables:

$$333 \quad U_{ni} = V_{ni} + \varepsilon_i \quad \forall i \text{ in } J \quad (2)$$

334 The systematic part of the utility function of individual  $n$  associated with the selected  
 335 alternative  $i$  is modelled as a linear function of the vector of the attributes  $x_i$  and associated  
 336 parameters  $\beta_n$ . If the unobserved error term  $\varepsilon_i$  is assumed to be i.i.d. extreme value type I,  
 337 the probability of individual  $n$  choosing alternative  $i$  out of  $J$  alternatives can be defined by  
 338 the MNL model:

$$339 \quad \pi_{ni} = \frac{\exp(\beta_n' x_i)}{\sum_{j=1}^J \exp(\beta_n' x_j)} \quad (3)$$

340 A property of the MNL model is the Independence of Irrelevant Alternatives (IIA). The IIA  
 341 property assumes that the choice probability of alternatives A and B is not influenced by  
 342 the addition or exclusion of any additional alternative in the choice set. In general, this is a  
 343 strong assumption that is often unrealistic. To relax this assumption, and to account for  
 344 taste heterogeneity across respondents, we estimated a LCM (Boxall and Adamowicz,  
 345 2002, Scarpa et al. 2003). The LCM endogenously and probabilistically assigns sampled  
 346 respondents to classes within which identical preferences are shared, but across which  
 347 preference differ. However, as these classes are latent (i.e. unobservable by the analyst) a  
 348 probabilistic equation explaining the probabilistic assignment of individual  $n$  into class  $C$   
 349 must be estimated. To specify the membership probability, we adopt a semi-parametric  
 350 form based on a class-specific constant term  $\alpha$  (Scarpa et al. 2003, Scarpa and Thiene,  
 351 2005), where for class 1 such term is set to zero for identification. Using a Logit  
 352 formulation for the class allocation model, the probability that individual  $n$  belongs to  
 353 segment  $C$  is given by (Bhat, 1997):

$$354 \quad \pi_{nc} = \frac{\exp(\alpha_c)}{\sum_{c=2}^C \exp(\alpha_c)}, \text{ where } \alpha_{c=1} = 0, \text{ for identification purposes.} \quad (4)$$

355 Given membership to class  $c$ , choice probabilities follow the random utility framework. The  
 356 probability that individual  $n$  chooses alternative  $i$ , conditional on belonging to taste group  $c$ ,  
 357 takes the logit form:

$$\pi_{ni|c} = \frac{\exp(\beta'_{nc} \mathbf{x}_i)}{\sum_{j=1}^J \exp(\beta'_{nc} \mathbf{x}_j)} \quad (5)$$

where  $\mathbf{x}_j$  represents the vector of attribute levels associated with alternative  $j$  and  $\beta_{nc}$  is a conformable vector of coefficients for class  $c$ .

WTP values for each attribute  $x$  in each class  $c$  are computed as the opposite of the ratio between the attribute coefficient  $\beta_{xc}$  and the price coefficient  $\beta_{COSTc}$ :

$$WTP_{xc} = - \frac{\beta_{xc}}{\beta_{COSTc}} \quad (6)$$

364

### 3.2 Benefit transfer

Benefit transfer can be conducted with different methods. One of the most common is to estimate a benefit function (Loomis 1992; Rosenberger and Loomis, 2003; Leon-Gonzales and Scarpa, 2008), through which a conditional estimate of the expected benefit can be derived. The simplest form of BFT uses an estimated function from a single primary study to calculate a calibrated welfare estimate for the policy site. This is often denoted as single-site BFT (Rolfe and Bennett, 2006; Johnston and Rosenberger, 2010). The benefit function can be expressed as:

$$y_{pk} = (\mathbf{x}_{pk}, \beta_p) \quad (7)$$

375

where  $y_{pk}$  is a predicted welfare estimate (in our case the WTP estimates from discrete choice models for a given ES change),  $\mathbf{x}_{pk}$  is a vector of determinants upon which one can condition the welfare estimate from change in ES  $p$  for people at site  $k$ , and  $\beta_p$  is the associated vector of coefficients. In our study, the elements in  $\mathbf{x}_{pk}$  were selected starting from some 60 different candidate determinants (see the Appendix for a full list) covering three main groups:

- 13 socio-demographic variables profiling respondents. Data were collected during the survey. The individual variables actually used in the vector  $\mathbf{x}_{pk}$  of the benefit transfer regression were age, sex, number of household members, occupation and average yearly income;
- 28 socio-demographic variables profiling the human dimension of municipalities sourced from official statistics (i.e. from the National Institute for Statistics, Istat). The conditioning variables included in the vector  $\mathbf{x}_{pk}$  of the benefit function were: total population, number of buildings, area covered by residential buildings, inhabitants' education and occupation, and population density;

- 14 territorial variables profiling the geographic dimension of municipalities and obtained via elaboration of the geographical layers of the territorial database of the Lombardy region. Among those variables, we focused on those that were most likely to influence the perceived value of the ES, such as (for example) the presence of sites that can be considered as substitutes of the two parks object of the study. The layers analyzed were those relative to urban parks and green areas, regional parks, land cover (Corine 2000), and scenic itineraries. From those layers, two types of data were obtained: the (logarithmic) distance of each municipality from substitute sites (as from ArcGIS 'near' function) and the coverage of substitute sites within each single municipality (as from ArcGIS 'intersect' function). The distance from each municipality to Ticino and Adamello Regional Parks was also used as a variable in the BFT, as it is known to influence the perceived value of the ES (e.g. land cover, log distance from the two Parks, etc.).

The resulting dataset was used to estimate the BFT by means of multiple linear regressions, using the software R. Statistical performance was tested for every candidate BFT determinant and only those with predictive power were maintained in the final specification used for the value transfer. The final BFT function estimated on the municipalities for which we had estimates were then used to predict the BFT for those sites  $k$ , using the generic BFT:

$$\widehat{WTP}_{pk} = \sum_{m=1}^{M(p)} (\hat{\alpha}_p + \hat{\beta}_{pm(p)}' \mathbf{x}_{km(p)}) \quad (8)$$

where  $\widehat{WTP}_{pk}$  is the predicted average WTP for the improvement of the ES  $p$  in municipality  $k$ ,  $\mathbf{x}_{km(p)}$  are the benefit determinants with values specific to the  $k$  municipality and  $\hat{\beta}_{pm(p)}$  is the generic estimated coefficient determinant for attribute  $p$  as obtained from the regressions run on the municipalities for which we had data. The total number of determinants acting as predictor for each  $p^{\text{th}}$  ES varies, as indicated by  $M(p)$ .

In order to transfer the values to non-sampled areas, we included the values of the variables for each municipality and we multiplied them with the associated estimated coefficients. The coefficients estimated for the first block of variables (that is socio-demographic variables profiling respondents), were associated with the average values of the municipality. For example, the coefficient associated with respondents' age was

multiplied with the average age of the municipality inhabitants. The average WTP estimate for each non-sampled municipality was then computed by adding up each term in the function. Finally, we aggregated the values at municipality level, by multiplying the average WTP estimate with the number of inhabitants. Aggregate results were also mapped to visualize their distribution across the region.

## **4. Results**

The main results from the research steps for the two DCEs with five BTF each are presented in the paragraphs below.

### **4.1 The choice experiments**

Results for the two DCEs are reported separately for the two conservation areas. We deal with each in turn.

#### **4.1.1 Preference for ES at Adamello Regional Park**

The total number of respondents to the questionnaire was 1,502, 39.7% of whom had visited the park. 97.3% of them (i.e. 1,461) completed the survey, 53.5% of whom revealed to be in favour of paying a regional tax to fund the park (53.5%), while 43.8% opposed the tax but nonetheless stated to value the benefits from these conservation areas (i.e. score > 2 on a Likert-scale). Only 2.7% of total respondents (i.e. 41 individuals) were classified as genuine protest respondents. The resulting choice models (MNL and LCM) were estimated on a panel of 17,532 choices (i.e. 1,461 respondents x 12 choice sets) with results reported in Table 5 showing that estimated marginal willingness-to-pay ( $WTP_m$ ) grows as the scope of the conservation policy intensifies with significant differences across ES types.  $WTP_m$  values are higher for meadow flora conservation (up to €8.19/year for 300 ha of managed meadow areas) and slope protection (up to €4.43/year for 45 km of safe road network). Increasing the number of wild fauna sighting sites is less valued: it ranges between €0.91/year and €1.07/year for additional five and eight sites, respectively. Low  $WTP_m$  values were observed also in the case of new floristic trails:  $WTP_m$  for an additional trail is €1.09/year, while  $WTP_m$  for three and five additional trails is €0.76/year and €1.93/year, respectively.

The specification search for the LCM identified eight to be the best number of classes fitting the observed data according to the Bayesian and to the Corrected Akaike's information criteria, BIC and CAIC, respectively (Table 3). We obtained a total of 127

parameters estimates, few of which are statistically insignificant. The resulting LCM gives significant or near-significant cost coefficient estimates, all with the expected negative sign, except for class 7. Most of the ES effects are significant across classes, but for three of the eight classes (3, 7 and 8) the ES changes show mostly insignificant effects on utility (Table 6).

We observe preference differences across the eight classes, as one would expect. Class 1 (26.6%) tends to prefer the ES of slope stability but there are no clear differences with the other attributes. In addition to slope stability, Class two (21.3%) is sensitive to flora conservation within meadow habitats close to the forest margin, but does not display a high WTP for the maintenance/restoration of dry-stone walls to enhance landscape value. Class three (15.9%) shows only two negative and significant coefficients: flora conservation over a 300 ha area (CON\_300) and the building of five additional floristic trails (FLOR\_6). This group seems to be interested in conservation and recreation aspects dealing with vegetation within the Park. Class four (12.1%) displays insignificant effects for dry stone-walls and low levels of slope stability, while all other attributes are significant and imply high  $WTP_m$  values, especially for flora conservation within meadow habitats close to the forest margin and floristic trails. To these respondents ES from flowers and plants matter.

Class five (8.6%) also cares about plants conservation, although with lower  $WTP_m$  values. Class six (8.6%) displays high  $WTP_m$  for policy on flora conservation when large areas are involved (at least 300 ha) and shows the highest  $WTP_m$  for the restoration of dry-stone walls among all the eight classes. Class seven has a low membership probability (only 3.5%) and is characterised by a positive COST coefficient. Individuals in this class seem generally uninterested in improving the current provisioning of ES, as suggested by the many negative WTP values. Finally, the residual Class eight, has smallest membership probability (3.2%), with a primary interest in slope stability and flora conservation within meadow habitats close to forest margins. This class displays a very low marginal value of money and has very large  $WTP_m$ ; it might be a class of wealthy respondents or a group of respondents with strong preferences for these attributes.

#### **4.1.2 Preference for ES at Ticino Regional Park**

The survey for these ES was administered to 1,500 respondents, 50.8% of whom visited the park, with only 2.9% of observations (43 individuals) dropped for protest voting. 53.3% of respondents stated to be in favour of paying a regional tax to improve the park area.



492 The MNL model considered a total number of 17,484 observations (i.e. 1,457 respondents  
 493 x 12 choice sets). As reported in Table 7, all but two coefficient estimates are insignificant,  
 494 the rest show positive  $WTP_m$ , with the highest  $WTP_m$  estimates identified for carbon  
 495 sequestration: respondents are willing to pay from €2.77/year for 5% emission reductions  
 496 to €9.61/year for 20% reduction. Positive  $WTP_m$  are estimated also for Ticino river water  
 497 quality (€0.58/year for 1 additional indicator species and €1.55/year for two species), water  
 498 meadow conservation (€0.89/year for the conservation of additional 80 ha and €1.18/year  
 499 for additional 130 ha), and scenic views with screened detractors (€0.87/year, €0.56/year  
 500 and €1.43/year for additional 6, 8 and 12 screened detractors, respectively).  
 501 In the LCM, the specification search for the DCE data on ES produced by Ticino RP  
 502 showed substantial heterogeneity, with information criteria preferring a 7-class model and  
 503 97 parameters estimates (Table 4). Most utility coefficient estimates are significant for all  
 504 classes, except for classes two (21.4%) and four (about 13.8%). Class four is also the only  
 505 class showing a positive COST coefficient (Table 8).  
 506 Class one (21.9%) includes people who appreciate all attributes except for (a) thematic  
 507 trails that might have been considered to be already supplied at the appropriate level, and  
 508 they are insignificantly different from zero; and (b) low improvements on water quality. The  
 509 highest  $WTP_m$  estimate values are observed for CO<sub>2</sub> emission reduction (RCO\_20) and  
 510 landscape (BVED\_12). Class two (21.4%) is focused only on high levels of CO<sub>2</sub> emission  
 511 reduction.  
 512 A similar pattern of preferences are found in Class three (16.5%) that shows much higher  
 513  $WTP_m$  values for CO<sub>2</sub> emission reduction, which are significant at all levels, compared to  
 514 Class two. Class four (13.8%) displays an unexpected positive value for the COST  
 515 coefficient, which prevents us from computing meaningful WTP estimates. Class five  
 516 (11.5%) shows an unusual pattern of alternating coefficient signs. Similarly to classes two  
 517 and three, there is a clear preference for strong CO<sub>2</sub> reduction and some WTP for high  
 518 levels of scenic views. Class six (11.4%) shows a very low marginal utility of money and  
 519 consequently high levels of  $WTP_m$  values, possibly for the same reasons as classe eight in  
 520 the sample for Adamello RP. What emerges in this class is a strong preference for CO<sub>2</sub>  
 521 reduction, which increases as the effort to reduce it increases, as one would expect. It also  
 522 shows substantive interest in other ES, but only when high policy effort is made. Finally,  
 523 class seven is the one with smallest membership probability (3.5%).  
 524  
 525

### 526 **4.1.3 Mapping**

527 Individual-specific  $WTP_m$  values were computed using the panel LCM estimator, for all ES  
528 of the two DCE surveys. Mean and standard deviations of the distributions of the values  
529 are reported in appendix. These values were averaged across each municipality and used  
530 for a preliminary mapping. Maps have been developed for all ES and their policy levels.  
531 For the purpose of illustration we report here only some selected results: those for slope  
532 stability (Adamello RP) and carbon sequestration (Ticino RP) in Figures 3 and 4,  
533 respectively. Average WTPs for slope stability across all districts range between €2.28 for  
534 35 km (STAB\_35) and €7.64 for 45 km (STAB\_45) of safe road network (baseline: 10 km).  
535 In general, municipalities with higher population densities tend to be associated with  
536 higher WTP values for the stability of slopes compared to low-density ones. While WTP  
537 values for the 45 km level are positive for most of the municipalities, the number of  
538 municipalities showing negative WTP values is much higher for the 35 km level, which  
539 tends to demonstrate that current provision is deemed highly insufficient.  
540 Average WTP values for reduced CO<sub>2</sub> emissions range between €8.30 for 5% reduction  
541 and €24.75 for 20% reduction (baseline: 0%). The geographical distribution of averaged  
542  $WTP_m$  values for a 5% reduction in CO<sub>2</sub> shows that these are positive for almost all  
543 municipalities, with many of them (mostly in Milan area and in the central part of the  
544 region) ranking over €7.50/person (all estimates are annual local tax payments for a period  
545 of five years). This is even more evident when considering a 20% reduction level,  
546 especially in the Central-Southern part of the region, but also in municipalities within and  
547 close-to the Ticino RP. Carbon sequestration seems to be perceived as a relevant ES by  
548 population throughout the region and for any attribute level.

549

### 550 **4.2 Benefit transfer**

551 The estimation of implied individual non-market benefits from the selected ES can be  
552 obtained only for a sub-set of the municipalities of Lombardy: those that were sampled  
553 (study municipalities). However, with adequate data and the determinants of such values,  
554 separate benefit function transfers for each ES were estimated and used to infer predicted  
555 values for all other “policy” municipalities. Of course, this process is tentative and has no  
556 intention to be policy-prescriptive, but only illustrative. This extended the “guesstimate” of  
557 average values across the entire 1,523 municipalities in Lombardy, multiplied by their  
558 respective population. As described in section 3.2, the first step to predict values in non-  
559 sampled municipality was the estimation of the benefit transfer function. As an example, in

Table 9 we report the results of the estimation of the linear regression for the ES floristic trails (level creation of 2 additional trails) in the Adamello RP. After testing the predicting power of all variables (see Appendix) we choose as our final model a specification, which includes only coefficients statistically significant at the 80% level ( $p < 0.2$ ). Literature suggests that this is an acceptable threshold for statistical significance in benefit transfer studies (Rosenberg and Loomis, 2000). Education (*edu*) has a positive effect on average WTP, whereas age (*age*) has a negative effect, suggesting that older and less educated individuals perceive less benefit from the improvement of this ES. Total population of the municipality (*ln\_pop\_tot*) has a positive effect on WTP. As highly urbanized cities usually offer scarce ES, it appears reasonable that residents of those areas would benefit from natural areas service improvement as they can easily visit them. The percentage of municipality soil covered by sparse vegetation (*s\_sparse*) has a positive effect on average WTP, whereas the logarithm of the distance from the Adamello Park (*l\_dist*) has a negative sign. This appears plausible, as individuals living far from the park are likely to perceive less benefit for the improvement of its ES, as suggested by the vast literature on distance decay (e.g. Schaafsma et al., 2013). Similar results, in terms of variables with significant effect and coefficient signs were obtained for the other two levels of the floristic trails attribute.

By using the coefficients estimated through the linear regression and the values of the related variables in each municipality, we then estimated the average WTP in each policy site. Finally, we aggregated the data at municipality level by multiplying the average WTP with the number of inhabitants. Table 10 summarizes the aggregate results for the ES provided by floristic trails in the Adamello RP. The benefits estimated for additional trails were mostly below €10,000/municipality: 79.1% in the case of 2 additional trails, 75.7% in the case of 4 and 48.1% in the case of 6. Nevertheless, for 6 additional trails, 21.1% of municipalities showed a total WTP higher than €30,000 (Table 7 and Figure 5). These figures can be used by managers of protected areas and policy makers to support their choices in terms, for example, of budget allocation and investments, including grants and subsidies. While trails are seen as an important asset at regional scale, as confirmed by the fact that in 2017 the Regional Council passed a new law recognizing the Regional Trail Network to value local natural and cultural resources, investment in trail maintenance activities within the whole regional Natura 2000 network between 2008 and 2011 only averaged about €156.000 (Gatto et al., 2015), which is a far lower amount than the total WTP that might be derived through the benefit transfer exercise.

594

## 595 **5. Discussion, limitations and further research**

596 The economic rationale behind investing in protected natural areas, including Natura 2000  
597 sites, has given place to a participated debate in Europe (Hoyos *et al.*, 2012). Since  
598 management costs for the EU-wide Natura 2000 network are expected to increase  
599 (Gantioler *et al.*, 2014), motivating the financing of such investments represents a key  
600 political issue. Equitable and efficient taxation schemes must implement the beneficiary  
601 pays principle, and hence crucially depend on accurate estimates of the magnitudes of  
602 private benefits and their localization, as well as how their provision can be achieved by  
603 specific management policy actions implemented in a cost-effective manner.

604 With few exceptions, estimates from both MNL models and LCMs are consistent with the  
605 “more is better” principle: increases in ES supply correlate with increases in WTP value  
606 estimates. So, non-market values satisfy the theoretical validity criterion. Furthermore,  
607 LCMs accounts for how preferences vary across respondents. Such preference variation  
608 should be appropriately heeded by local policy-makers to spatially target the ES delivery  
609 as well as to equitably spread the associated tax burden.

610 The 6-steps methodological approach adopted for the study is not just instrumental to the  
611 research, but it represents one of its most valuable outputs. Yet, we are fully aware it  
612 would need substantive improvements to enhance evidence-based policy action, quality of  
613 research findings and, ultimately impacts. These would include:

- 614 • **FBT improvement:** it is recommendable to assess further WTP determinants by  
615 revising the list of socio-demographic and territorial variables used to develop the  
616 function(s) as well as to adopt spatial-econometric approaches, in order to take into  
617 consideration spatial correlation among data/WTP values. Although it is impossible  
618 to identify impacts of these measures *a priori*, it can be assumed they are likely to  
619 improve the quality of BT outputs. Testing spatial autocorrelation (e.g. by using  
620 specific functions available in many mapping tools) may also improve the FBT and  
621 further research in this area should be conducted;
- 622 • **Data enhancement:** stratifying the sample according to the distance from the two  
623 study sites (among other features) was a methodological prerequisite, but it was  
624 only partly achieved. This reflected on the quality of outcomes and should be  
625 carefully considered in future surveys. Although the two study sites are highly  
626 representative of mountainous and lowland areas in Lombardy, thus allowing a first  
627 approximation of the regional territory, they cannot cover the full range of situations

and nuances characterising the whole Natura 2000 network at the regional scale. It is then recommendable to perform additional surveys and studies at the scale of single sites or groups of sites in order to enrich data in both qualitative and quantitative terms. As an additional issue, some of the benefits from recreational ES are accrued to visitors from outside the region, i.e., the analysis presented within this paper, being focused on people resident in Lombardy, does not necessarily cover the entire population of beneficiaries for these ES;

- **Systematic data collection and management:** following on from the previous point, it would be worthwhile to develop a systematic data collection at the regional scale. This could consist of a dataset to be collected and reported according to standardised methodologies (units, frequency, periods, geo-reference, etc.) at the appropriate scale (e.g. single Natura 2000 site or cluster of sites) and with reference to one or more well-defined ES. Specific checklists could be developed and provided to site managers to fill them. Further data management could allow the identification of panels worth collecting in representative points so as to account for location-specific effects over time.

## 6. Conclusions

Conservation areas can be managed to produce different levels of flows of valuable ES in the form of local public goods, which are valued by residents in a manner that varies across the land and according to individual preferences. This poses a challenge to raising funds to finance such policies in an optimal manner. This paper makes a first empirical attempt at dealing with the issue of economic valuation of ES generated from two areas of conservation within the Natura 2000 network of Lombardy, Italy. Making local public good beneficiaries pay for ES requires a clear understanding of the relationship between policy actions and distribution of WTP over the land, and hence over different jurisdictions. Our proposed methodology can, in principle, deliver such information with the required degree of accuracy. A full mapping of partly measured and partly inferred estimates of marginal WTP were obtained for all municipal districts using a system of benefit function transfers. These estimates, once validated, could represent a base onto which elaborate an efficient local public revenue system for ES, reflective of both, patterns of human settlement and ES benefits.

Besides providing some preliminary economic values, the research contributes to the development of a methodology for assessing and monitoring ES over time by mapping

662 and valuing them. Through further development and implementation of this methodology,  
663 regular monitoring and assessment of Natura 2000 benefits could be achieved and the  
664 database expanded in a cost-efficient manner. This would be in line with the requirements  
665 recently set by the environmental norms included within the 2015 (Italian) National Budget  
666 Law and - in more general terms - could provide an informative basis for developing future  
667 policies as well as supporting decision-making by other relevant actors (companies,  
668 citizens, private donors, etc.) in order to sustain the contribution of Natura 2000 areas to  
669 rural development and bio-based economy.

670

## 671 **Acknowledgments**

672 Authors would like to express their gratitude to Life+ GESTIRE Project for funding this  
673 research, and to ERSAF and Lombardy Region staff for providing technical and  
674 administrative support. Authors would also like to thank Pragma for online interviews and  
675 staff at Adamello and Ticino Regional Parks for providing valuable insights and inputs to  
676 set up questionnaires.

## 677 **References**

- 678 Bastian, O., 2013. The role of biodiversity in supporting ecosystem services in Natura  
679 2000 sites. *Ecol. Indic.* 24, 12-22.
- 680 Bateman, I.J., Day, B.H., Georgiou, S., Lake, I., 2006. The aggregation of environmental  
681 benefit values: Welfare measures, distance decay and total WTP. *Ecol. Econ.* 60,  
682 450-460.
- 683 Bateman, I.J., Harwood, A.R., Mace, G.M., Watson, R.T., Abson, D.J., Andrews, B.,  
684 Binner, A., Crowe, A., Day, B.H., Dugdale, S., Fezzi, C., Foden, J., Hadley, D.,  
685 Haines-Young, R., Hulme, M., Kontoleon, A., Lovett, A.A., Munday, P., Pascual, U.,  
686 Paterson, J., Perino, G., Sen, A., Siriwardena, G., van Soest, D., Termansen, M. ,  
687 2013. Bringing ecosystem services into economic decision-making: land use in the  
688 United Kingdom, *Science* 341,45-50.
- 689 Ben-Akiva, M., Walker, J., Bernardino, A.T., Gopinath, D.A., Morikawa, T.,  
690 Polydoropoulou, A., 1997. Integration of choice and latent variable models. Paper  
691 presented at the 1997 IATBR, University of Texas, Austin. Available at:  
692 [www.joanwalker.com/uploads/3/6/9/5/3695513/benakivawalkeretal\\_iclv\\_chapter\\_200](http://www.joanwalker.com/uploads/3/6/9/5/3695513/benakivawalkeretal_iclv_chapter_2002.pdf)  
693 [2.pdf](http://www.joanwalker.com/uploads/3/6/9/5/3695513/benakivawalkeretal_iclv_chapter_2002.pdf) [Last access: 19<sup>th</sup> September 2018].
- 694 Bergstrom, J.C., De Civita, P., 1999. Status of benefits transfer in the United States and  
695 Canada: a review. *Can. J. Agric. Econ.* 47, 79-87.
- 696 Bhat, C., 1997. An endogenous segmentation mode choice model with an application to  
697 intercity travel. *Transport. Sci.*, 31(1), 34-48.
- 698 Boxall, P.C., Adamowicz, W.L., Swait, J., Williams, M., Louviere, J.J., 1996. A comparison  
699 of stated preference methods for environmental valuation. *Ecol. Econ.* 18, 243-253.
- 700 Boxall, P.C., Adamowicz, W.L., 2002. Understanding heterogeneous preferences in  
701 random utility models: a latent class approach. *Environ. Resour. Econ.* 234, 421-446.
- 702 Campbell, D., Erdem, S., 2018. Including Opt-Out Options in Discrete Choice  
703 Experiments: Issues to Consider. *Patient.* 2018 Aug 2. doi: 10.1007/s40271-018-  
704 0324-6.
- 705 Campbell, D., Hutchinson, W.G., Scarpa, R., 2009. Using Choice Experiments to Explore  
706 the Spatial Distribution of Willingness to Pay for Rural Landscape Improvements.  
707 *Environ. Plann. A*, 41, 97-111.
- 708 Carlsson, F., Mørkbak, M.R., Olsen, S.B. 2012. The first time is the hardest: A test of  
709 ordering effects in choice experiments *Journal of Choice Modelling*, 5(2):19-37.
- 710 Caussade, S., Ortúzar, J.D., Rizzi, L.I., Hensher, D.A., 2005. Assessing the influence of  
711 design dimensions on stated choice experiment estimates *Transp. Res. Part B:*  
712 *Methodol.*, 39 (2005), 621-640.
- 713 Choicemetrics, 2014. Ngene v. 1.1.2, User manual and Reference Guide. Available at:  
714 <http://www.choice-metrics.com/download.html> [Last access: 19<sup>th</sup> September 2018].

- 715 CICES, 2017. Common International Classification of Ecosystem Services (CICES)  
716 Version 4.3. Available at: <http://cices.eu> [Last access: 19<sup>th</sup> September 2018]
- 717 COM, 2011. Final communication from the commission to the European Parliament, the  
718 Council, the Economic and Social Committee and the Committee of the Regions: our  
719 life insurance, our natural capital: an EU biodiversity strategy to 2020. European  
720 Commission, Brussels, May 3, 2011.
- 721 Chuan-Zhong, L., Kuuluvainen, J., Pouta, E., Rekola, M., Tahvonen, O., 2004. Using  
722 choice experiments to value the Natura 2000 conservation programs in Finland.  
723 Environ. Resour. Econ. 29, 361-374.
- 724 Christie, M., Rayment, M., 2012. An economic assessment of the ecosystem service  
725 benefits derived from the SSSI biodiversity conservation policy in England and  
726 Wales. Ecosyst. Ser. 1 (1), 70-84.
- 727 CTS, 2014. Piano di Comunicazione Progetto LIFE11NAT/IT/044 GESTIRE. Development  
728 of the strategy to manage the Natura 2000 network in the Lombardia Region.  
729 Available at:  
730 [http://www.pdc.minambiente.it/sites/default/files/progetti/gestire\\_piano\\_di\\_comunicazione.pdf](http://www.pdc.minambiente.it/sites/default/files/progetti/gestire_piano_di_comunicazione.pdf)  
731 [Last access: 19<sup>th</sup> September 2018]
- 732 Czajkowski, M., Giergiczny, M., Greene, W.H. 2014. Learning and fatigue effects  
733 revisited: Investigating the effects of accounting for unobservable preference and  
734 scale heterogeneity. Land Economics, 90(2):324-351.
- 735 Czajkowski, M., Budziński, W., Campbell, D., Giergiczny, M., Hanley, N., 2016. Spatial  
736 Heterogeneity of Willingness to Pay for Forest Management. Environ. Resource  
737 Econ. 68(3), 705-727.
- 738 Downing, M., Ozuna, T., 1996. Testing the reliability of the benefit transfer approach. J.  
739 Environ. Econ. Manage. 30, 316-322.
- 740 Eastwood, A., Brooker, R., Irvine, R.J, Artz, R.R.E., Norton, L.R., Bullock, J.M., Ross, L.,  
741 Fielding, D., Ramsay, S., Roberts, J., Anderson, W., Dugan, D., Cooksley, S.,  
742 Pakeman, R.J., 2016. Does nature conservation enhance ecosystem services  
743 delivery? Ecosyst. Ser. 17, 152-162.
- 744 Eurobarometer, 2010. Attitudes of Europeans towards the issue of biodiversity. Wave 2.  
745 Flash EB Series#290. Available at:  
746 [http://ec.europa.eu/public\\_opinion/flash/fl\\_290\\_en.pdf](http://ec.europa.eu/public_opinion/flash/fl_290_en.pdf) [Last access: 19<sup>th</sup> September  
747 2018]
- 748 European Commission, 2017. Natura 2000 Barometer. Available at:  
749 [http://ec.europa.eu/environment/nature/natura2000/barometer/index\\_en.htm](http://ec.europa.eu/environment/nature/natura2000/barometer/index_en.htm) [Last  
750 access: 19<sup>th</sup> September 2018]
- 751 Freeman III, A.M., Herriges, J.A., Kling, C.L., 2014. The measurement of Environmental  
752 and resource values: theory and methods. Routledge, London, 460 pp.



- 753 Gantioler, S., Rayment, M., ten Brink, P., McConville, A., Kettunen, M., Bassi, S., 2014.  
754 The costs and socio-economic benefits associated with the Natura 2000 network. *J.*  
755 *Sustainable Soc.* 6, 135-157.
- 756 Gatto, P., Maso, D., Leonardi, A., 2015. Analisi dei costi per la gestione della Rete Natura  
757 2000 in Regione Lombardia nel periodo 2008-2011 e di stima dei costi nel periodo  
758 2014-2020. Etifor, Padova. Available at: [http://www.naturachevale.it/wp-](http://www.naturachevale.it/wp-content/uploads/2016/08/allegato-VIII_C21.pdf)  
759 [content/uploads/2016/08/allegato-VIII\\_C21.pdf](http://www.naturachevale.it/wp-content/uploads/2016/08/allegato-VIII_C21.pdf) [Last access: 19<sup>th</sup> September 2018]
- 760 Gibson, H., Hanley, N., Wright, R., 2004. An Economic Assessment of the Costs and  
761 Benefits of Natura 2000 Sites in Scotland. Scottish Executive 2004. Environment  
762 Group Research Report 2004/05. Available at:  
763 [www.gov.scot/Resource/Doc/47251/0014580.pdf](http://www.gov.scot/Resource/Doc/47251/0014580.pdf) [Last access: 19<sup>th</sup> September  
764 2018]
- 765 Gómez-Baggethun, E., Ruiz-Pérez, M., 2011. Economic valuation and the  
766 commodification of ecosystem services. *Prog. Phys. Geogr.* 35, 613-628.
- 767 Greene, W.H., Hensher, D.A., Rose, J.M., 2005. Using classical simulation-based  
768 estimators to estimate individual WTP values. In: Scarpa, R., Alberini, A. (Eds).  
769 Applications of simulation methods in environmental and resource economics (Vol.  
770 6). Springer Science and Business Media, Dordrecht, pp. 17-33.
- 771 Hagenaars, J.A., McCutcheon, A.L., 2002. Applied Latent Class Analysis. Kluwer,  
772 Dordrecht, 480 pp.
- 773 Harrison, P.A., Vandewalle, M., Sykes, M.T., Berry, P.M., Bugter, R., de Bello, F., Feld,  
774 C.K., Grandin, U., Harrington, R., Haslett, J.R., Jongman, R.H.G., Luck, G.W.,  
775 Martins da Silva, P., Moora, M., Settele, J., Sousa, J.P., Zobel, M., 2010. Identifying  
776 and prioritising services in European terrestrial and freshwater ecosystems.  
777 *Biodivers. Conserv.* 19, 2791-2821.
- 778 Heal, G.M., Barbier, E.B., Boyle, K.J., Covich, A.P., Gloss, S.P., Hershner, C.H. *et al.*,  
779 2005. Valuing Ecosystem Services: Toward Better Environmental Decision-Making.  
780 Committee on Assessing and Valuing the Services of Aquatic and Related Terrestrial  
781 Ecosystems - Water Science and Technology Board - Division on Earth and Life  
782 Studies. The National Academies Press, Washington D.C., 291 pp.
- 783 Hensher, D.A., 2006. How do respondents process stated choice experiments? Attribute  
784 consideration under varying information load. *J. Appl. Econom.*, 21, 861-878.
- 785 Holland, B.M., Johnston, R.J., 2017. Optimized quantity-within-distance models of spatial  
786 welfare heterogeneity. *J. Environ. Econ. Manag.* 85, 110-129.
- 787 Hoyos, D., Mariel, P., Pascual, U., Etxano, I., 2012. Valuing a Natura 2000 network site to  
788 inform land use options using a discrete choice experiment: an illustration from the  
789 Basque Country. *J. For. Econ.* 18 (4), 329-344.
- 790 Johnston, R.J., Duke, J.M., 2009. Willingness to Pay for Land Preservation across States  
791 and Jurisdictional Scale: Implications for Benefit Transfer. *Land Econ.* 85, 217-237.

- 792 Johnston, R.J., Moeltner, K., 2014. Meta-Modeling and Benefit Transfer: The Empirical  
793 Relevance of Source-Consistency in Welfare Measures. *Environ. Resour. Econ.*  
794 59(3), 337-361.
- 795 Johnston, R.J., Ramachandran, M., 2013. Modeling Spatial Patchiness and Hot Spots in  
796 Stated Preference Willingness to Pay. *Environ. Resource Econ.* 59, 363-  
797 387. Johnston, R.J., Rosenberger, R., 2010. Methods, trends and controversy in  
798 contemporary benefit transfer. *J. Econ. Surv.*, 24(3), 479-510.
- 799 Johnston, R.J., Jarvis, D., Wallmo, K., Lew, D.K., 2015. Multiscale Spatial Pattern in  
800 Nonuse Willingness to Pay: Applications to Threatened and Endangered Marine  
801 Species. *Land Econ.* 93, 739-761.
- 802 Johnston, R.J., Rolfe, J., Rosenberger, R.S., Brouwer, R. (Eds), 2015. *Benefit Transfer of  
803 Environmental and Resource Values: A Guide for Researchers and Practitioners.*  
804 Springer: Dordrecht, 582 pp.
- 805 Kamakura, W.A., Wedel, M., 2004. An empirical Bayes procedure for improving individual-  
806 level estimates and predictions from finite mixtures of multinomial logit models. *J.*  
807 *Bus. Econ. Stat.* 22, 121-125.
- 808 Kirchhoff, S., Colby, B.G., LaFrance, J.T., 1997. Evaluating the performance of benefit  
809 transfer: An empirical inquiry. *J. Environ. Econ. Manage.* 33 (1), 75-93.
- 810 Ispra, 2015. Il consumo di suolo in Italia. Edizione 2015. Rapporti 218/2015. Ispra, Istituto  
811 Superiore per la Protezione e la Ricerca Ambientale, Roma.
- 812 Istat, 2015. Noi Italia. 100 statistiche per capire il Paese in cui viviamo. Available at:  
813 <http://noi-italia.istat.it> [Last access: 19<sup>th</sup> September 2018]
- 814 Leon-Gonzalez, R. and Scarpa, R. 2008. Improving multi-site benefit functions via  
815 Bayesian model averaging: A new approach to benefit transfer. *J. of Envir. Econ. and*  
816 *Manag.*, 56(1), 50-68.
- 817 Loomis, J.B., 1992. The evolution of a more rigorous approach to benefit transfer: benefit  
818 function transfer. *Water Resour. Res.*, 28(3), 701-705.
- 819 Loomis, J., Roach, B., Ward, F., Ready, R., 1995. Testing transferability of recreation  
820 demand models across regions - A study of corps of engineer reservoirs. *Water*  
821 *Resour. Res.* 31(3), 721-730.
- 822 Louviere, J.J., 1991. Experimental choice analysis: Introduction and overview. *J. Bus. Res.*  
823 23, 291-297.
- 824 Luce, R.D., 1959. *Individual choice behavior. A theoretical analysis.* Wiley: New York, 151  
825 pp.
- 826 Maes, J., Egoh, B., Willemen, L., Liqueste, C., Vihervaara, P., Schägner, J.P., Grizzetti, B.,  
827 Drakou, E.G., La Notte, A., Zulian, G., Bouraoui, F., Paracchini, M.L., Braatd, L.,  
828 Bidoglio, G., 2012. Mapping ecosystem services for policy support and decision  
829 making in the European Union. *Ecosyst. Ser.* 1(1), 31-39.

830 McFadden, D., 1973. Conditional logit analysis of qualitative choice behavior. In Bateman  
831 *et al.* (2002). Economic valuation with stated preference techniques: A manual.  
832 Edward Elgar, Northampton, pp. 105-142.

833 McFadden, D., 1986. The choice theory approach to market research. *Mark. Sc.* 5, 275-  
834 297.

835 MELS, 2013. Italy's fifth national report to the Convention on Biological Diversity. Italian  
836 Ministry for the Environment, Land and Sea. Available at: [www.cbd.int/doc/world/it/it-](http://www.cbd.int/doc/world/it/it-nr-05-en.pdf)  
837 [nr-05-en.pdf](http://www.cbd.int/doc/world/it/it-nr-05-en.pdf) [Last access: 19<sup>th</sup> September 2018]

838 Moeltner, K., Boyle, K.J., Paterson, R.W., 2007. Meta-analysis and benefit transfer for  
839 resource valuation-addressing classical challenges with Bayesian modelling. *J.*  
840 *Environ. Econ. Manage.* 53 (2), 250-269.

841 Moeltner, K., Rosenberger, R.S., 2014. Cross-context benefit transfer: a Bayesian search  
842 for information pools. *Am. J. Agric. Econ.* 96 (2), 469-488.

843 Morey, E., Thiene M., 2017. Can personality traits explain where and with whom you  
844 recreate? A latent-class site-choice model informed by estimates from a mixed-mode  
845 LC cluster models with latent-personality traits, *Ecol. Econ.* 138, 223-237.

846 Pagiola, S., von Ritter, K., Bishop, J., 2004. Assessing the economic value of ecosystem  
847 conservation. The World Bank Environment Department - Paper n. 101. The World  
848 Bank, Washington D.C., 58 pp.

849 Pettenella, D., Thiene, M., Scarpa, R., Mattea, S., Masiero, M., Franceschinis, C., Comini,  
850 B., Cavalli, G., Gagliazzi, E., Fracassi, G., Spinelli, O., Bellisari, L., Zaghi, D., Rampa,  
851 A., 2016. Stima del valore socio-economico della rete Natura 2000 in Lombardia  
852 Azione A5. Rapporto finale. LIFE+11 NAT/IT/044 "Development of the Strategy to  
853 manage the Nature 2000 network in the Lombardia Region" GESTIRE. Available at:  
854 [http://www.naturachevale.it/gestire/wp-content/uploads/2014/04/Stima-del-valore-](http://www.naturachevale.it/gestire/wp-content/uploads/2014/04/Stima-del-valore-socio-economico-della-Rete-Natura-2000-in-Lombardia.pdf)  
855 [socio-economico-della-Rete-Natura-2000-in-Lombardia.pdf](http://www.naturachevale.it/gestire/wp-content/uploads/2014/04/Stima-del-valore-socio-economico-della-Rete-Natura-2000-in-Lombardia.pdf) [Last access: 19<sup>th</sup>  
856 September 2018]

857 Pirard, R., 2012. Market-based instruments for biodiversity and ecosystem services: a  
858 lexicon. *Env. Sci. Pol.* 19-20: 59-68.

859 Popescu, V.D., Rozylowicz, L., Niculae, I.M., Cucu, A.L., Hartel, T., 2014. Species,  
860 habitats, society: an evaluation of research supporting EU's Natura 2000 network.  
861 *PLoS ONE* 9(11): e113648. doi:10.1371/journal.pone.0113648.

862 Popper, R., Kroll, J., Magidson, J., 2004. Application of latent class models to food product  
863 development: A case study. *Sawthooth Conference Proceedings*: 89-112.

864 Regione Lombardia, 2018. Osservatorio Regionale della Biodiversità. Il monitoraggio  
865 scientifico regionale. Programma di monitoraggio 2017-2018.  
866 [http://www.biodiversita.lombardia.it/sito/index.php?option=com\\_content&view=article](http://www.biodiversita.lombardia.it/sito/index.php?option=com_content&view=article&id=107:monitoraggio-scientifico&catid=79&Itemid=464)  
867 [&id=107:monitoraggio-scientifico&catid=79&Itemid=464](http://www.biodiversita.lombardia.it/sito/index.php?option=com_content&view=article&id=107:monitoraggio-scientifico&catid=79&Itemid=464) [Last access: 19<sup>th</sup>  
868 September 2018]

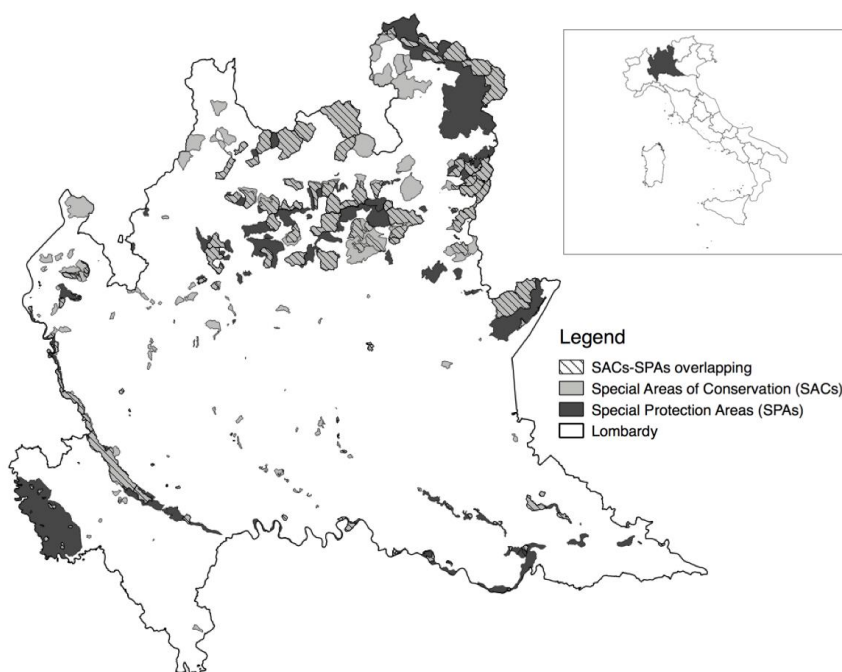
- 869 Rolfe, J., Bennett, J. (Eds), 2006. Choice Modelling and the Transfer of Environmental  
870 Values. Edward Elgar: Cheltenham, 272 pp.
- 871 Rosenberger, R.S., Loomis, J.B., 2003. Benefit transfer. In: Champ, P.A., Boyle, K.J.,  
872 Brown, T.C. (Eds). A primer on non-market valuation. Springer: Dordrecht, pp. 445-  
873 482.
- 874 Ruckelshaus, M., McKenzie, E., Tallis, H., Guerry, A., Daily, G., Kareiva, P., Polasky, S.,  
875 Ricketts, T., Bhagabati, N., Wood, S.A., Bernhardt, J., 2015. Notes from the field:  
876 lessons learned from using ecosystem service approaches to inform real-world  
877 decisions. *Ecol. Econ.* 115, 11-21.
- 878 Sagebiel, J., 2017. Preference heterogeneity in energy discrete choice experiments: A  
879 review on methods for model selection. *Renew. Sust. Energ. Rev.* 69, 804-811.
- 880 Sarrias, M.A and Daziano R.A., 2018. Individual-Specific Point and Interval Conditional  
881 Estimates of Latent Class Logit Parameters. *J. of Choice Mod.* 27:50-61.
- 882 Scarpa, R., Chilton, S.M., Hutchinson, W.G., Buongiorno, J., 2000. Valuing the  
883 recreational benefits from the creation of nature reserves in Irish forests. *Ecol. Econ.*  
884 33(2), 237-250.
- 885 Scarpa, R., Thiene, M., 2005. Destination choice models for rock climbing in the  
886 Northeastern Alps: a latent-class approach based on intensity of preference. *Land*  
887 *Econ.* 81 (3), 426-444.
- 888 Scarpa, R., Campbell, D. Hutchinson, W.G., 2007. Benefit estimates for landscape  
889 improvements: sequential Bayesian design and respondents' rationality in a choice  
890 experiment study. *Land Econ.* 83(4), 617-634.
- 891 Scarpa, R., Thiene, M., Marangon, F., 2008. Using flexible taste distributions to value  
892 collective reputation for environmentally-friendly production methods. *Can. J. Agric.*  
893 *Econ.* 56, 145-162.
- 894 Scarpa, R., Drucker, A. Anderson, S. Ferraes-Ehuan, N. Risopatron, V. G. C. R. and  
895 Rubio-Leonel, O. 2003. Valuing Animal Genetic Resources in Peasant Economies:  
896 The Case of the Box Keken Creole Pig in Yucatan, *Ecol. Econ.*, vol. 45(3):427-443.
- 897 Schaafsma, M., Brouwer, R., Gilbert, A., van der Bergh, J., Wagtendonk, A., 2013.  
898 Estimation of Distance-Decay Functions to Account for Substitution and Spatial  
899 Heterogeneity in Stated Preference Research. *Land Econ.* 89, 514-537.
- 900 Schägner, J.P., Brander, L., Maes, J., Hartje, V., 2013. Mapping ecosystem services'  
901 values: Current practice and future prospects. *Ecosyst. Ser.* 4, 33-46.
- 902 Schirpke, U., Scolozzi, R., De Marco, C., 2013. Analisi dei servizi ecosistemici nei siti  
903 pilota. Parte 4: Selezione dei servizi ecosistemici. Report del progetto Making Good  
904 Natura (LIFE+11 ENV/IT/000168), EURAC research, Bolzano, 42 pp.
- 905 Schirpke, U., Scolozzi, R., De Marco, C., Tappeiner, U., 2014. Mapping beneficiaries of  
906 ecosystem services flows from Natura 2000 sites. *Ecosyst. Ser.* 9, 170-179.

- 907 Schirpke, U., Marino, D., Marucci, A., Palmieri, M., Scolozzi, R., 2017. Operationalising  
908 ecosystem services for effective management of protected areas: Experiences and  
909 challenges. *Ecosyst. Ser.* 28, 105-114.
- 910 Schirpke, U, Scolozzi, R., Da Re, R., Masiero, M., Pellegrino, D., Marino, D., 2018.  
911 Recreational ecosystem services in protected areas: a survey of visitors to Natura  
912 2000 sites in Italy. *JORT* 21, 39-50.
- 913 Schulp, C., Burkhard, B., Maes, J., Van Vliet, J., Verburg, P.H., 2014. Uncertainties in  
914 ecosystem service maps: a comparison on the European scale. *PLoS ONE* 9(10):  
915 e109643. doi:10.1371/journal.pone.0109643.
- 916 Silvis, H.J., van der Heide, C.M., 2013. Economic viewpoints on ecosystem services.  
917 Statutory Research Tasks Unit for Nature and the Environment (WOT Natuur &  
918 Milieu) – Wageningen University, Den Haag, 70 pp.
- 919 Smith V.K., 1993. Nonmarket Valuation of Environmental Resources: An Interpretive  
920 Appraisal. *Land Econ.* 69(1), 1-26.
- 921 Smith, V.K., Van Houtven, G., Pattanayak, S.K., 2002. Benefit Transfer via Preference  
922 Calibration: “Prudential Algebra” for Policy. *Land Econ.* 78(1), 132-152.
- 923 ten Brink, P., Bassi, S., Badura, T., Gantioler, S., Kettunen, M., Mazza, L., Hart, K., 2013.  
924 The Economic benefits of the Natura 2000 Network. Final Sythesis Report. Available  
925 at: [http://ec.europa.eu/environment/nature/natura2000/financing/docs/ENV-12-018\\_LR\\_Final1.pdf](http://ec.europa.eu/environment/nature/natura2000/financing/docs/ENV-12-018_LR_Final1.pdf) [Last access: 19<sup>th</sup> September 2018]  
926
- 927 TEEB, 2013. Guidance Manual for TEEB Country Studies. Version 1.0. Available at:  
928 [http://www.teebweb.org/media/2013/10/TEEB\\_GuidanceManual\\_2013\\_1.0.pdf](http://www.teebweb.org/media/2013/10/TEEB_GuidanceManual_2013_1.0.pdf) [Last  
929 access: 19<sup>th</sup> September 2018]
- 930 Thiene, M., Galletto, L., Scarpa, R., Boatto, V., 2013. Determinants of WTP for Prosecco  
931 wine: a latent class regression with attitudinal responses. *Br. F. J.* 115(2), 279- 299.
- 932 Thiene M., Scarpa R., Louviere J., 2015. Addressing preference heterogeneity, multiple  
933 scales and attribute attendance with a correlated finite mixing model of tap water  
934 choice, *Environ. Resour. Econ.* 62(3), 637-656.
- 935 Train, K., 2003. *Discrete Choice Methods with Simulation*. Cambridge University Press,  
936 Cambridge, 378 pp.
- 937 Turbé, A., De Toni, A., Benito, P., Lavelle, P., Ruiz, N., Van der Putten, W.H., Labouze, E.,  
938 Mudgal S., 2010. Soil biodiversity: functions, threats and tools for policy makers. Bio  
939 Intelligence Service, IRD, and NIOO, Technical Report European Commission (DG  
940 Environment) 2010-049. Available at:  
941 [http://ec.europa.eu/environment/soil/pdf/biodiversity\\_report.pdf](http://ec.europa.eu/environment/soil/pdf/biodiversity_report.pdf) [Last access: 19<sup>th</sup>  
942 September 2018]
- 943 UNEP-WCMC, 2004. Species Data. World Conservation Monitoring Centre of the United  
944 Nations Environment Programme. Available at: <http://www.unep-wcmc.org> [Last  
945 access: 19<sup>th</sup> September 2018]

- 946 Vázquez-Polo, F.J., Guerra, N., León, C., Riera, P., 2002. A Bayesian model for benefit  
947 transfer: application to national parks in Spain. *Appl. Econ.* 34, 749-757.
- 948 Veldwijk, J., Lambooij, M.S., de Bekker-Grob, E.W., Smit H.A., de Wit G.A., 2014. The  
949 effect of including an opt-out option in discrete choice experiments. *PLoS One*. 2014  
950 Nov 3;9(11):e111805. doi: 10.1371/journal.pone.0111805.
- 951 Vossler, C.A., Doyon, M., Rondeau, D., 2012. Truth in consequentiality: theory and field  
952 evidence on discrete choice experiments, *Am. Econ. J.: Microeconomics* 4 (4), 145-  
953 171.
- 954 Wünscher, T., Engel, S., Wunder, S., 2008. Spatial targeting of payments for  
955 environmental services: A tool for boosting conservation benefits. *Ecol. Econ.* 4, 822-  
956 833.
- 957 Ziv, G., Hassall, C., Bartkowski, B., Cord, A.F., Kaim, A., Kalamandeen, M., Landaverde-  
958 González, P., Melo, J.L.B., Seppelt, R., Shannon, C., Václavík, T., Zoderer, B.M.,  
959 Beckmann, B.M., 2017. A bird's eye view over ecosystem services in Natura 2000  
960 sites across Europe. *Ecosyst. Ser.* 30, Part B, 287-298.
- 961 Zulian, G., Stange, E., Woods, H., Carvalho, L., Dick, J., Andrews, C., Baró, F., Vizcaino,  
962 P., Barton, D.N., Nowel, M., Rusch, G.M., Autunes, P., Fernandes, J., Ferraz, D.,  
963 Ferreira dos Santos, R., Aszalós, R., Arany, I., Czúcz, B., Priess, J.A., Bürger-  
964 Patricio, G., Lapola, D., Mederly, P., 2018. Practical application of spatial ecosystem  
965 service models to aid decision support. *Ecosyst. Ser.* 29, 465-480.

966 **Figures and tables**

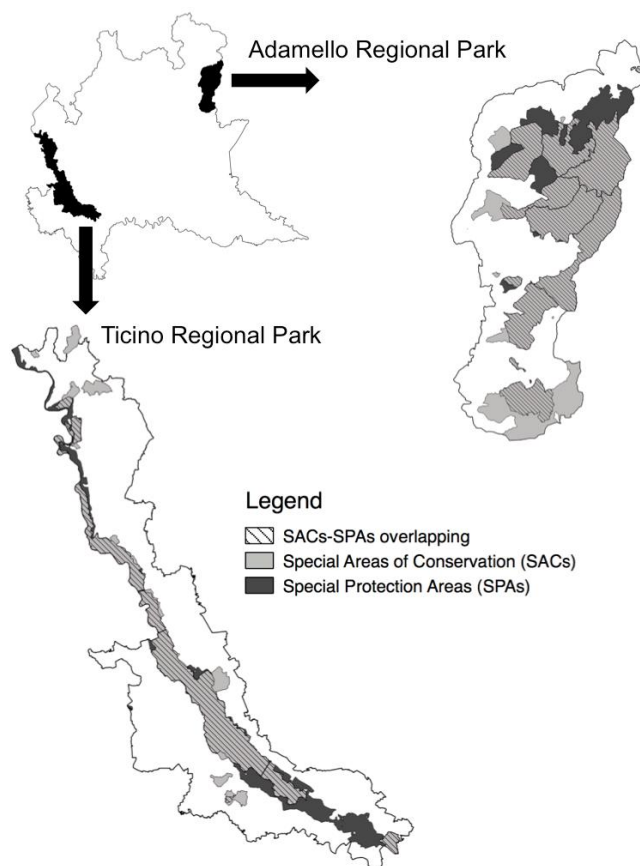
967



968

969 **Figure 1: SACs and SPAs in Lombardy**

970



971

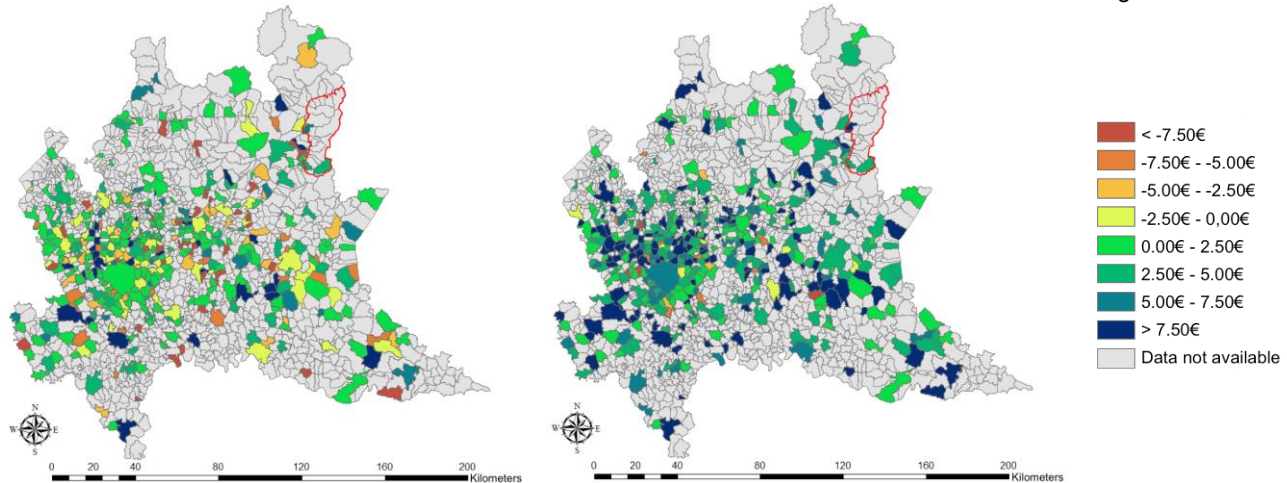
972 **Figure 2: Study areas – Adamello and Ticino RPs**



a. 35 km safe road network

b. 45 km safe road network

Legend

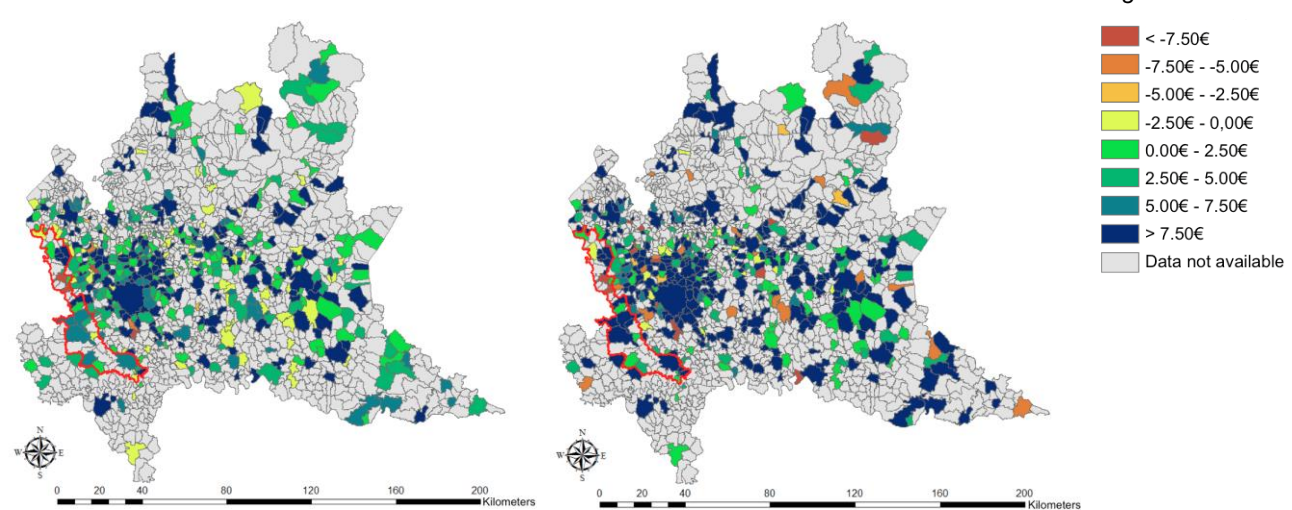


**Figure 3:** Distribution of average WTP (€) within Lombardy municipalities covered by the survey for 2 different levels of the attribute "slope stability" in Adamello RP

a. 5% CO<sub>2</sub> reduction

b. 20% CO<sub>2</sub> reduction

Legend



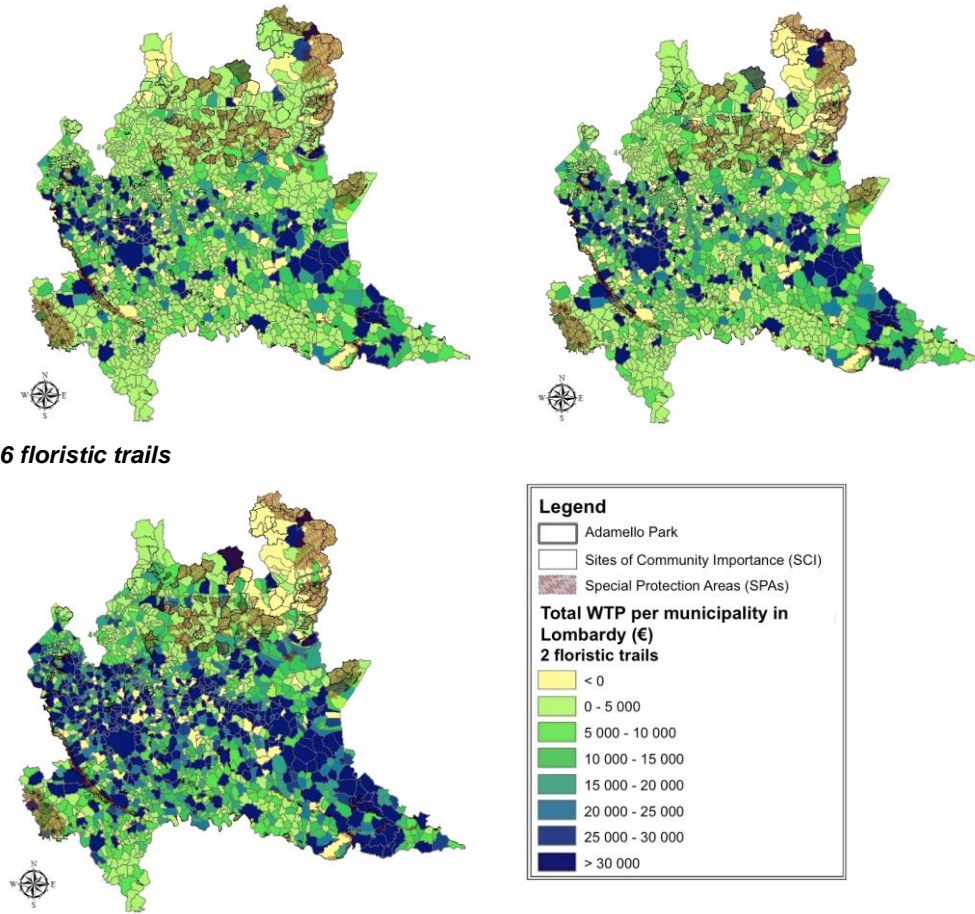
**Figure 4:** Distribution of average WTP (€) within Lombardy municipalities covered by the survey for 2 different levels of the attribute "carbon sequestration" in Ticino RP



2 floristic trails

4 floristic trails

6 floristic trails



**Figure 5:** Total WTP at municipal scale for different levels of the attribute "Floristic trails" in the Adamello RP

990

Attributes	Abbreviations	Levels
<b>Slope Stability</b> Increased slope stability and consequent road safety	STAB_10	10 km safe road network (1/6 on 60 km) (baseline)
	STAB_20	20 km safe road network (1/3 on 60 km)
	STAB_35	35 km safe road network (7/12 on 60 km)
	STAB_45	45 km safe roads (9/12 on 60 km)
<b>Flora Conservation</b> Management practices to conserve flora in particular at the forest/meadow interface	CON_0	0 ha meadows managed (baseline)
	CON_200	200 ha meadows managed (1/16 of total meadow area)
	CON_250	250 ha meadows managed (1/13 of total meadow area)
	CON_300	300 ha meadows managed (1/11 of total meadow area)
<b>Fauna</b>  Presence of fauna sighting sites	FAUN_2	2 fauna sighting sites (baseline)
	FAUN_5	5 fauna sighting sites (+3 sites)
	FAUN_7	7 fauna sighting sites (+5 sites)
	FAUN_10	10 fauna sighting sites (+8 sites)
<b>Recreation</b>  Development of new trails to valorize floristic features in the area	FLOR_1	1 floristic trail (baseline)
	FLOR_2	2 floristic trails (+1 trail)
	FLOR_4	4 floristic trails (+3 trails)
	FLOR_6	6 floristic trails (+5 trails)
<b>Landscape</b>  Maintenance of dry-stone wall as an indicator of landscape quality	SEC_450	450 ha dry-stone wall in good state (baseline)
	SEC_453	453 ha dry-stone wall in good state (+3 ha)
	SEC_455	455 ha ha dry-stone wall in good state (+5 ha)
<b>Tax</b>	COST	Regional tax (0€, 2€,5€,10€,15€,20€)

991 **Table 1: Adamello RP: Attributes, their abbreviations and levels**

992

Attributes	Abbreviations	Levels
<b>Carbon sequestration</b>  Improved carbon sequestration through appropriate management practices	RCO_0	0% CO <sub>2</sub> emission reduction (baseline)
	RCO_5	5% CO <sub>2</sub> emission reduction (-0,42 tCO <sub>2</sub> /year/inhabitant)
	RCO_10	10% CO <sub>2</sub> emission reduction (-0,84 tCO <sub>2</sub> / year/inhabitant)
	RCO_20	20% CO <sub>2</sub> emission reduction (-1,67 tCO <sub>2</sub> / year/inhabitant)
<b>Water quality</b>  Number of fish species that indicate actual improvement of Ticino river water quality	WATQ_2	Ticino River water quality (2 indicator species) (baseline)
	WATQ_3	Ticino River water quality (3 indicator species) (+1 species)
	WATQ_4	Ticino River water quality (4 indicator species) (+2 species)
<b>Biodiversity</b>  Conservation of valuable floral features, in particular water meadows	MAR_320	320 ha water meadow (baseline)
	MAR_400	400 ha water meadow (+80ha managed meadows)
	MAR_450	450 ha water meadow (+130ha managed meadows)
<b>Landscape</b>  Number of scenic views with screened detractors as an indicator of landscape quality	BVED_0	0 scenic views with screened detractors (0 on 25) (baseline)
	BVED_6	6 scenic views with screened detractors (1/4 of total detractors)
	BVED_8	8 scenic views with screened detractors (1/3 of total detractors)
	BVED_12	12 scenic views with screened detractors (1/2 of total detractors)
<b>Recreation</b>  Development of new thematic trails to offer additional recreational opportunities in the area	ITIN_62	62 thematic trails (baseline)
	ITIN_65	65 thematic trails (+3 trails)
	ITIN_67	67 thematic trails (+5 trails)
<b>Tax</b>	COST	Regional tax (0€, 2€,5€,10€,15€,20€)

993 **Table 2: Ticino RP: Attributes, their abbreviations and levels**

994

995

Number of classes	k	LL	BIC(LL)	AIC(LL)	AIC3(LL)	CAIC(LL)
1	15	-15017.3	30143.82	30064.51	30079.51	30158.82
2	31	-12917.5	26060.84	25896.94	25927.94	26091.84
3	47	-12448.5	25239.48	24991	25038	25286.48
4	63	-12261.4	24981.95	24648.88	24711.88	25044.95
5	79	-12097.8	24771.26	24353.6	24432.6	24850.26
6	95	-11999.2	24690.61	24188.36	24283.36	24785.61
7	111	-11915.1	24639.01	24052.16	24163.16	24750.01
8	127	-11847.1	24619.59	23948.16	24075.16	24746.59
9	143	-11794.4	24630.77	23874.75	24017.75	24773.77

**Table 3: Adamello RP: Information criteria**

Number of classes	k	LL	BIC(LL)	AIC(LL)	AIC3(LL)	CAIC(LL)
1	13	-15253	30600.78	30532.09	30545.09	30613.78
2	27	-13326.6	26849.96	26707.29	26734.29	26876.96
3	41	-12719.3	25737.15	25520.5	25561.5	25778.15
4	55	-12277.7	24956.09	24665.46	24720.46	25011.09
5	69	-12076.7	24656.08	24291.47	24360.47	24725.08
6	83	-11910.8	24426.27	23987.69	24070.69	24509.27
7	97	-11839.6	24385.85	23873.29	23970.29	24482.85
8	111	-11783.5	24375.56	23789.03	23900.03	24486.56

**Table 4: Ticino RP: Information criteria**

Choice (see Table 1)	Coefficient	Std. Err.	z	95% Confidence Interval		WTP <sub>m</sub>	Significance
COST	-0.108	0.002	-71.31	-0.111	-0.105		***
STAB_20	0.125	0.027	4.67	0.072	0.177	1.156	***
STAB_35	0.305	0.027	11.32	0.252	0.358	2.828	***
STAB_45	0.478	0.026	18.54	0.428	0.529	4.433	***
CON_200	0.621	0.028	22.21	0.566	0.676	5.751	***
CON_250	0.693	0.029	24.30	0.637	0.748	6.417	***
CON_300	0.884	0.027	33.31	0.832	0.936	8.193	***
FAUN_5	0.015	0.028	0.54	-0.039	0.069	0.137	
FAUN_7	0.098	0.027	3.66	0.046	0.151	0.912	***
FAUN_10	0.116	0.026	4.48	0.065	0.166	1.071	***
FLOR_2	0.118	0.027	4.33	0.065	0.171	1.093	***
FLOR_4	0.082	0.027	3.00	0.028	0.135	0.758	***
FLOR_6	0.201	0.026	8.12	0.158	0.258	1.925	***
SEC_453	0.001	0.022	0.04	-0.042	0.044	0.009	
SEC_455	0.001	0.022	0.43	-0.033	0.052	0.087	

Note: \*\*\*, \*\*, \* = 99%, 95%, 90% significance

**Table 5: Adamello RP: MNL estimates**

Choice (see Table 1)	Class 1			Class 2			Class 3			Class 4			Class 5			Class 6			Class 7			Class 8		
	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>
COST	-1.20	-9.45		-0.22	-15.91		-0.01	-1.80		-0.05	-5.97		-0.37	-6.35		-0.09	-8.11		0.14	7.04		-0.03	-1.91	
STAB_20	-0.18	-0.82	-0.15	-0.02	-0.22	-0.09	0.10	1.51	9.09	0.11	1.02	2.20	0.78	2.28	2.11	-0.17	-1.39	-1.89	0.15	0.77	-1.07	2.25	2.91	75.00
STAB_35	0.57	3.07	0.48	0.13	1.28	0.59	-0.14	-1.78	-12.73	0.30	2.52	6.00	1.45	2.67	3.92	-0.57	-3.18	-6.33	0.06	0.28	-0.43	3.28	4.86	109.33
STAB_45	0.96	5.29	0.80	0.65	6.76	2.95	0.08	0.95	7.27	0.88	6.47	17.60	1.48	4.05	4.00	0.47	2.64	5.22	0.76	4.02	-5.43	4.66	5.98	-55.33
CON_200	0.69	2.88	0.58	0.33	3.17	1.50	-0.01	-0.19	-0.91	2.33	8.44	46.60	5.56	4.10	15.03	-0.27	-1.25	-3.00	0.37	1.66	-2.64	1.39	4.85	46.33
CON_250	0.72	4.02	0.60	0.46	4.17	2.09	0.06	0.86	5.45	2.87	10.20	57.40	6.01	4.15	16.24	-0.53	-2.91	-5.89	0.44	2.19	-3.14	1.10	3.15	36.67
CON_300	0.69	4.16	0.58	0.82	7.75	3.73	0.20	2.50	18.18	3.83	12.56	76.60	6.86	4.62	18.54	1.19	8.37	13.22	0.88	4.49	-6.29	1.88	6.28	62.67
FAUN_5	-0.13	-0.68	-0.11	0.16	1.63	0.73	0.03	0.42	2.73	0.39	3.18	7.80	0.34	0.95	0.92	-0.42	-3.06	-4.67	0.10	0.56	-0.71	0.30	1.11	10.00
FAUN_7	0.48	2.49	0.40	0.44	4.63	2.00	0.11	1.41	10.00	0.28	2.17	5.60	0.09	0.32	0.24	-0.21	-1.33	-2.33	-0.06	-0.36	0.43	0.29	1.06	9.67
FAUN_10	0.40	2.09	0.33	0.41	4.42	1.86	0.06	0.75	5.45	0.58	4.59	11.60	0.47	1.60	1.27	0.34	1.95	3.78	0.11	0.55	-0.79	0.27	1.08	9.00
FLOR_2	0.06	0.23	0.05	0.10	1.18	0.45	0.11	1.50	10.00	0.39	3.72	7.80	0.28	1.12	0.76	-0.19	-1.39	-2.11	-0.11	-0.60	0.79	-0.41	-1.35	-13.67
FLOR_4	0.02	0.10	0.02	0.37	4.14	1.68	0.13	1.85	11.82	0.50	4.60	10.00	0.28	1.28	0.76	-0.66	-4.50	-7.33	0.05	0.24	-0.36	-0.36	-0.75	-12.00
FLOR_6	0.61	3.18	0.51	0.47	5.49	2.14	0.32	4.13	29.09	0.87	8.06	17.40	0.41	1.54	1.11	-0.19	-1.22	-2.11	0.42	2.17	-3.00	-0.22	-0.43	-7.33
SEC_453	-0.04	-0.26	-0.03	-0.27	-3.67	-1.23	-0.10	-1.75	9.09	-0.08	-0.84	-1.60	0.10	0.45	0.27	0.30	2.02	3.33	0.00	-0.01	0.00	0.07	0.24	2.33
SEC_455	-0.03	-0.16	-0.03	-0.06	-0.74	-0.27	0.04	0.72	3.64	-0.07	-0.80	-1.40	0.32	1.53	0.86	0.58	4.14	6.44	-0.01	-0.06	0.07	-0.10	-0.44	-3.33
Log-likelihood -11,847.08																								
Size	26.64			21.33			15.97			12.07			8.60			8.59			3.54			3.26		

**Table 6: Adamello RP: LCM estimates (coefficients statistically significant at 90% level in bold)**

Choice (see Table 2)	Coefficient	Std. Err.	z	95% Confidence Interval		WTP <sub>m</sub>	Significance
COST	-0.109	0.002	-66.91	-0.112	-0.106		***
RCO_5	0.302	0.028	10.76	0.247	0.357	2.773	***
RCO_10	0.592	0.028	20.87	0.537	0.648	5.434	***
RCO_20	1.047	0.029	36.00	0.990	1.104	9.606	***
WATQ_3	0.063	0.021	2.95	0.021	0.106	0.582	***
WATQ_4	0.169	0.022	7.80	0.127	0.212	1.552	***
MAR_400	0.097	0.022	4.37	0.054	0.141	0.890	***
MAR_450	0.128	0.022	5.97	0.086	0.171	1.178	***
BVED_6	0.095	0.026	3.68	0.044	0.145	0.868	***
BVED_8	0.062	0.03	2.06	0.003	0.120	0.564	***
BVED_12	0.156	0.027	5.83	0.104	0.209	1.434	***
ITIN_65	-0.228	0.022	-1.04	-0.066	0.020	-0.209	
ITIN_67	-0.034	0.021	-1.61	-0.076	0.007	-0.316	

Note: \*\*\*, \*\*, \* = 99%, 95%, 90% significance

**Table 7: Ticino RP: MNL model estimates**

Choice (see Table 2)	Class 1			Class 2			Class 3			Class 4			Class 5			Class 6			Class 7		
	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>	Coeff.	z	WTP <sub>m</sub>
COST	<b>-0.20</b>	-18.88		<b>-1.68</b>	-9.32		<b>-0.23</b>	-15.81		<b>0.02</b>	3.47		<b>-0.10</b>	-9.88		<b>-0.02</b>	-2.90		<b>-1.84</b>	-3.76	
RCO_5	<b>0.25</b>	2.88	1.25	<b>-0.56</b>	-2.15	-0.33	<b>2.53</b>	11.86	11.00	-0.06	-0.81	3.00	<b>0.27</b>	2.25	2.70	<b>1.49</b>	5.52	74.50	<b>22.66</b>	3.31	12.32
RCO_10	<b>0.33</b>	3.47	1.65	0.13	0.48	0.08	<b>3.00</b>	13.36	13.04	0.00	-0.01	0.00	-0.23	-1.54	-2.30	<b>3.28</b>	12.26	164.00	<b>9.35</b>	2.73	5.08
RCO_20	<b>0.58</b>	4.88	2.90	<b>1.76</b>	4.57	1.05	<b>4.54</b>	15.74	19.74	0.08	0.96	-4.00	<b>2.07</b>	10.14	20.70	<b>5.07</b>	16.52	253.50	<b>53.59</b>	3.58	29.13
WATQ_3	0.03	0.40	0.15	0.28	1.21	0.17	<b>0.17</b>	1.87	0.74	0.09	1.53	-4.50	<b>-0.64</b>	-5.80	-6.40	<b>0.20</b>	1.69	10.00	<b>-6.14</b>	-3.76	-3.34
WATQ_4	<b>0.20</b>	2.96	1.00	<b>0.43</b>	1.79	0.26	<b>0.49</b>	5.46	2.13	<b>0.29</b>	4.71	-14.50	<b>-0.56</b>	-5.15	-5.60	<b>0.33</b>	2.79	16.50	<b>-2.51</b>	-2.59	-1.36
MAR_400	<b>0.20</b>	3.01	1.00	0.28	1.14	0.17	0.05	0.58	0.22	-0.08	-1.43	4.00	-0.12	-1.35	-1.20	0.13	1.14	6.50	<b>-6.14</b>	-3.11	-3.34
MAR_450	<b>0.16</b>	2.57	0.80	0.17	0.89	0.10	0.01	0.11	0.04	-0.01	-0.12	0.50	<b>0.27</b>	2.99	2.70	0.43	3.88	21.50	-0.16	-0.20	-0.09
BVED_6	<b>0.41</b>	5.01	2.05	-0.34	-1.50	-0.20	-0.06	-0.58	-0.26	<b>0.13</b>	1.77	-6.50	-0.05	-0.37	-0.50	0.47	3.51	23.50	1.47	1.48	0.80
BVED_8	<b>0.28</b>	2.91	1.40	<b>-0.96</b>	-2.54	-0.57	<b>-0.45</b>	-3.39	-1.96	<b>0.26</b>	3.21	-13.00	<b>-0.85</b>	-4.68	-8.50	0.18	1.36	9.00	<b>-7.56</b>	-4.08	-4.11
BVED_12	<b>0.57</b>	5.80	2.85	<b>0.52</b>	1.73	0.31	0.03	0.24	0.13	<b>0.27</b>	3.58	-13.50	<b>1.22</b>	8.82	12.20	<b>0.33</b>	2.30	16.50	<b>25.42</b>	3.48	13.82
ITIN_65	-0.06	-0.86	-0.30	-0.15	-0.66	-0.09	<b>-0.23</b>	-2.44	-1.00	-0.05	-0.83	2.50	<b>-0.47</b>	-5.09	-4.70	<b>-0.25</b>	-2.13	-12.50	<b>-2.78</b>	-2.67	-1.51
ITIN_67	0.08	1.23	0.40	0.03	0.17	0.02	-0.06	-0.71	-0.26	-0.01	-0.23	0.50	-0.03	-0.36	-0.30	-0.19	-1.22	-3.50	<b>0.42</b>	2.17	2.14
Log-likelihood	-11,839.64																				
Size	21.96			21.41			16.49			13.77			11.43			11.42			3.52		

**Table 8: Ticino RP: LCM estimates (coefficients statistically significant at 90% level in bold)**

Variable	Estimate	t	Pr(> t )
Intercept	6.79	2.64	0.008281
edu	0.7	1.53	0.127536
age	-0.1	3.42	0.000644
ln(pop_tot)	0.01	1.53	0.12615
s_sparse	3.44	1.57	0.117078
ln_dist	-0.83	1.75	0.079082
edu x ln(pop_tot)	0.02	-2.01	0.04547

Adjusted R-squared: 0.2232

Multiple R-squared: 0.2304

F-statistic: 18.24 on 7 and 1460 DF

p-value: <0.001

**Table 9: Ticino RP: estimates of the BTF for 2 new floristic trails**

Total WTP (€)	2 floristic trails		4 floristic trails		6 floristic trails	
	no. municipalities	% on total	no. municipalities	% on total	no. municipalities	% on total
Less than 0	81	5.2	110	7.1	89	5.8
0 – 5,000	863	55.9	748	48.4	383	24.8
5,001 – 10,000	277	17.9	311	20.1	271	17.6
10,001 – 15,000	118	7.6	138	8.9	177	11.5
15,001 – 20,000	50	3.2	68	4.4	137	8.9
20,001 – 25,000	24	1.6	27	1.7	89	5.8
25,001 – 30,000	12	0.8	17	1.1	73	4.7
More than 30,000	119	7.7	125	8.1	325	21.0
<b>Total</b>	<b>1,544</b>	<b>100.0</b>	<b>1,544</b>	<b>100.0</b>	<b>1,544</b>	<b>100.0</b>

1015 **Table 10:** Distribution of Lombardy municipalities within different total WTP classes for different levels of the  
1016 attribute “Floristic trails” in the Adamello RP (absolute and % values)

1017

1018

## 1019 Appendix

1020 List of variables used for the multiple regression

Acronym	Description	Unit
Age	Age	Years
Male	Male	Dummy
Female	Female	Dummy
Edu	Education level	Ordinal
Members	Number of household members	Units
I_self	Job-status: self-employed	Dummy
I_empl	Job-status: employed	Dummy
L_frla	Job-status: free lance	Dummy
L_entrep	Job-status: entrepreneur	Dummy
I_stud	Job-status: student	Dummy
I_hw_re	Job-status: housewife or househusband/retired	Dummy
I_une	Job-status: unemployed	Dummy
Income	Annual income	€
Tot_pop	Total resident population	Units
av_fam	Average number of household members	Units
Num_hous	Number of residential houses within the municipality	Units
Area	Surface area of resident population's houses	km <sup>2</sup>
t_illit	Education level of resident population (> 6 years): illiterate	Percent over total municipal population above 6 years old
t_il_nq	Education level of resident population (> 6 years): illiterate, no educational qualification	Percent over total municipal population above 6 years old
t_prim	Education level of resident population (> 6 years): primary school	Percent over total municipal population above 6 years old
t_seco	Education level of resident population (> 6 years): secondary school	Percent over total municipal population above 6 years old
t_high	Education level of resident population (> 6 years): high-school	Percent over total municipal population above 6 years old
t_terz	Education level of resident population (> 6 years): tertiary education degree	Percent over total municipal population above 6 years old
t_univ	Education level of resident population (> 6 years): university degree	Percent over total municipal population above 6 years old
u_sec_19	High school diploma index by gender and age classes: males, over 19 years old	Percent over total municipal population above 19 years old
d_sec_19	High school diploma index by gender and age classes: females, over 19 years old	Percent over total municipal population above 19 years old
no_p_m	Early leavers index, 1 <sup>st</sup> school cycle, per gender: males	Percent over total municipal population
no_p_f	Early leavers index, 1 <sup>st</sup> school cycle, per gender: females	Percent over total municipal population
oc_agri	Employees by economic activity sector: agriculture	Percent over total employed population
oc_ind	Employees by economic activity sector: industry	Percent over total employed population
oc_com	Employees by economic activity sector: trade	Percent over total employed population
oc_fin	Employees by economic activity sector: finance	Percent over total employed population
oc_oth	Employees by economic activity sector: other activities	Percent over total employed population
cp_occ	Occupational status of resident population: employed	Percent over total municipal population
cp_look	Occupational status of resident population: looking for a job	Percent over total municipal population
cp_ret	Occupational status of resident population: retired	Percent over total municipal population
cp_stud	Occupational status of resident population: student	Percent over total municipal population
cp_house	Occupational status of resident population: housewife or househusband	Percent over total municipal population
cp_other	Occupational status of resident population: other	Percent over total municipal population
av_inc	Average income per municipality	€
Density	Population density	inhabitants/km <sup>2</sup>
d_road	Road distance municipality-Park	km
I_d_road	Logarithm of road-distance municipality-Park	Linear m
I_d_reg	Logarithm of distance from the closest Regional Park	Linear m

1021  
1022

Acronym	Description	Unit
l_d_urb	Logarithm of distance from the closest urban park and green area	Linear m
l_d_pan	Logarithm of distance from the closest scenic itineraries	Linear m
s_past	Area occupied by pastures	Percent over total municipal area
s_bush	Area occupied by bushes	Percent over total municipal area
s_sparse	Area occupied by sparse/scattered vegetation	Percent over total municipal area
s_urban	Area occupied by urban green areas	Percent over total municipal area
s_conif	Area occupied by coniferous forests	Percent over total municipal area
s_broadl	Area occupied by broadleaves forests	Percent over total municipal area
s_mixed	Area occupied by mixed forests	Percent over total municipal area
s_moorl	Area occupied by moorlands	Percent over total municipal area
s_mead	Area occupied by meadows	Percent over total municipal area

Statistics of the variables used in the computation of the BTF

Variable	Mean	Standard Deviation
<b>Individual mWTP in Adamello RP</b>		
mWTP stab_20	2.71	14.66
mWTP stab_35	2.28	20.38
mWTP stab_45	7.64	30.78
mWTP con_200	7.62	21.31
mWTP con_250	9.15	25.43
mWTP con_300	15.04	37.82
mWTP faun_5	1.05	4.35
mWTP faun_7	2.11	5.23
mWTP faun_10	2.92	6.42
mWTP flor_2	1.45	5.68
mWTP flor_4	1.62	7.47
mWTP flor_6	4.55	16.72
mWTP sec_453	-0.86	4.27
mWTP sec_455	0.72	3.36
<b>Individual mWTP in Ticino RP</b>		
mWTP rco_5	8.33	38.23
mWTP rco_10	13.48	88.21
mWTP rco_20	24.82	141.90
mWTP mWTP atq_3	-0.56	10.57
mWTP mWTP atq_4	-0.38	30.82
mWTP mar_400	0.84	7.60
mWTP mar_450	2.02	11.56
mWTP bved_6	1.33	21.49
mWTP bved_8	-2.08	24.83
mWTP bved_12	2.84	40.09
mWTP itin_65	-1.56	11.82
mWTP itin_67	-0.05	2.62
<b>Exogenous variables</b>		
edu	3.54	1.10
reddito	23.96	16.86
pop_tot	212649.32	439973.25
l_p_tot	10.24	1.99
med_fam	4.26	75.38
num_abit	106527.66	223432.06



l_n_abit	9.46	2.04
area	8515643.12	17537466.64
l_area	13.94	2.00
t_analf	0.63	1.34
t_an_nt	6.35	1.29
t_elem	19.74	3.98
t_medie	29.85	4.47
t_super	30.86	3.44
t_terz	0.49	1.20
t_univ	12.08	5.83
u_sec_19	41.06	24.95
d_sec_19	37.44	23.68
no_p_m	4.87	19.70
no_p_f	8.26	88.77
oc_agri	2.37	2.76
oc_ind	33.93	11.11
oc_com	18.89	2.70
oc_fin	17.58	6.77
oc_alt	27.23	6.22
cp_occ	39.27	2.79
cp_cer	3.01	1.10
cp_pens	20.35	2.93
cp_stud	5.08	0.97
cp_cas	7.03	1.70
cp_alt	3.02	1.18
cpm_occ	22.24	1.61
densita	2215.42	2241.69
d_strad	34012.67	33378.01
l_strad	9.34	2.73
l_d_reg	3.59	3.29
l_d_giar	4.10	3.33
l_d_pan	9.40	2.23
s_pasc	2.69	7.29
s_arb	1.04	3.06
s_rada	0.81	3.41
s_urban	0.09	0.92
s_conif	3.82	10.48
s_latif	16.12	22.92
s_misti	5.28	11.98
s_brugh	0.27	1.51
s_stabili	1.67	5.46
sop50	0.32	0.47
sot50	0.68	0.47

1023

1024

1025 Choice task example for Ticino RP

Attributes	Option A	Option B	Option C
------------	----------	----------	----------

1026

1027

CO <sub>2</sub> emission reductions (%)	-10%	-10%	-5%
Water quality	4 species	4 species	3 species
Water meadow conservation (ha)	450 hectares	400 hectares	320 hectares
Scenic views with screened detractors (n.)	8 sites	0 sites	6 sites
Thematic trails (n.)	65 trails	67 trails	62 trails
Annual Tax (for 5 years)	15 Euro	5 Euro	10 Euro

Choice task example for Adamello RP

1028

1029

1030

1031

1032

1033

1034

1035

1036

1037

1038

Online survey (example)

**TESAF** Dipartimento Territorio e Sistemi Agro-Forestali Università di Padova

**ERSAF** ENTE REGIONALE PER I SERVIZI ALL'AGRICOLTURA E ALLE FORESTE

**Regione Lombardia**

**Scenario 3 – Click on the option corresponding to the alternative you prefer**

	Option A	Option B	Option C
CO <sub>2</sub> emission reductions (%)	-10%	-10%	-5%
Water quality	4 species	4 species	3 species
Water meadow conservation (hectares)	450 hectares	400 hectares	320 hectares
Scenic views with screened detractors (n.)	8 sites	0 sites	6 sites
Thematic trails (n.)	65 trails	67 trails	62 trails
Annual Tax (for 5 years)	15 Euro	5 Euro	10 Euro

If you want to have a look to proposals for baseline improvements, please click "Help" button below

Back Continue Cancel Help Pause

1039